

## Durham E-Theses

# The effects of refinery effluent on the invertebrate intertidal fauna and flora of Little Wick Bay, Milford Haven 

Archer-Thomson, John Henry Stuart

## How to cite:

Archer-Thomson, John Henry Stuart (1979) The effects of refinery effluent on the invertebrate intertidal fauna and flora of Little Wick Bay, Milford Haven, Durham theses, Durham University. Available at Durham E-Theses Online: http://etheses.dur.ac.uk/9021/

## Use policy

The full-text may be used and/or reproduced, and given to third parties in any format or medium, without prior permission or charge, for personal research or study, educational, or not-for-profit purposes provided that:

- a full bibliographic reference is made to the original source
- a link is made to the metadata record in Durham E-Theses
- the full-text is not changed in any way

The full-text must not be sold in any format or medium without the formal permission of the copyright holders.
Please consult the full Durham E-Theses policy for further details.

## of Little Wick Bay, Milford Haven

## by JOHN HENRY STUART ARCHBR-THOMSON

The copyright of this thesis rests with the author.
No quotation from it should be published without his prior written consent and information derived from it should be acknowledged.

## Submitted for the Degree of Master of Science <br> (Advanced course in Ecology) <br> University of Durham <br> Department of Zoology <br> 1979



## of Little Wick Bay, Milford Haven

by ARCHER-THOMSON J.H.S.


#### Abstract

\section*{ABSTRACT}

The effects of continuous low level oil pollution from a refinery effluent, on intertidal fauna and flora in Little Wick Bay are investigated.

A shore survey of the intertidal species on six transects at varying distances from the effluent discharge point, is carried out and the results compared with past surveys of the same transects. Any differences or similarities in the findings are related to the environmental agencies in operation since the first survey.

A detailed investigation of the size classes and abundance of the Limpet Patella vulgata at each of the six transects and a quantitative analysis of Petroleum Oil Pollutants in P. vulgata by Infra red Spectrophotometry is carried out in an attempt to relate findings to the effluent discharge.


## ACKNOWLEDGEMENTS

I would like to express my sincere gratitude to all those people at O.P.R.U. who were so helpful with both advice and equipment, and say a special thank you to Saran Petpiroon for his help, and for allowing me to use his unpublished results from his Ph.D. thesis.

I would also like to thank the Chemistry Department of the University College of Swansea for their exceptional generosity with time, help and equipment without which the whole project would have been impossible. A special thank you is due to Professor H Purnell, Dr J Ballantine and especially to Jean Pierre Aubertin.

Finally, I woula like to thank ur $\boldsymbol{Y} \mathrm{k}$ evans or $\ln$ urnam university for his help and advice throughout the project and especially with the preparation of the manuscript.

Dedicated to Maggie, with thanks.

## TABLE OF CONTENTS

pageAbstract
Title Page
Acknowledgements
List of Tables and Illustrations
Introduction ..... 1
Methods ..... 10
I The Shore Survey ..... 10
II Studies of Patella vulgata at the six Little Wick Transects ..... 1.4
(i) Investigation of the sizes of P. vulgata in Little ..... 13
Wick Bay
(ii) Quantitative analysis of Petroleum oil Pollutants in ..... 15
P. vulgata by Infra red Spectrophotometry
(a) Sampling and Sample preparation ..... 16
(b) Extraction and Analysis ..... 17
Discussion ..... 24
The Shore Survey ..... 24
(i) The distribution of Littoral Animals and Plants in 1979 ..... 24
in relation to the refinery effluent discharge point
(ii) Possible Explanations of the observed Species ..... 28 Distributions
(iii) Comparisons of the present Survey with those of Past ..... 33 Surveys
Summary of the Shore Survey ..... 35
II Studies of Patella vulgata at the six Little Wick Transects ..... 36
(i) Comparisons of the Size-frequency distributions of ..... 36
P. vulgata at the six Little Wick Transects
Page
(ii) Quantitative Analysis of Petroleum Oil Pollutants in ..... 44
P. vulgata by infra red Spectrophotometry
Summary of the studies of P. vulgata ..... 49
Appendix ..... 52
A (1) Results for the 1979 Shore Survey (Archer-Thomson) ..... 53
A (2) Results for the 1978 Shore Survey (Petpiroon) ..... 60
A (3) Results for the 1970 Shore Survey (Crapp) ..... 67
A (4) Criteria of Abundance ..... 74
A (5) Exposure score sheet (Ballentine 1961) ..... 77
A (6) Raw Limpet Data ..... 79
 ..... 23
References ..... 94

## List of Tables and Illustrations

|  |  | Page ${ }^{\text {P }}$ |
| :---: | :---: | :---: |
| Figure One | South West Weles, showing Milford Haven and the |  |
|  | surrounding area |  |
| Figure Two | Showing the Position of Little Wick Bay near the | 6 |
|  | town of Milford Haven |  |
| Figure Three | The Little Wick Transects in relation to the Outflow | 6 |
|  | Pipe and Esso Jetty |  |
| Figure Four | Showing the Present model of the "Crosstaff" after | 12 |
|  | modification by Crapp in 1970 (Taken from Crapps |  |
|  |  |  |
| Figure Five | The sample area investigated at each station along | 12 |
|  | the Transect Tape |  |
| Figure Six | Showing sample points from the six Little Wick | 18 |
|  | Transects |  |
| Figure Seven | Detail of a single Sample Area for any given Iransect | 18 |
| Figure Eight | An idealised trace from a Spectrophotometer readout | 18 |
| Figure Ten | Approximate Shore Cross sections for the six Little | 39 |
|  | Wick Transects |  |
| Figure Nine | Distributions of six common species at the six | 25 |
|  | Iittle Wick Transects |  |
| Figure Eleven | Size Frequency Distributions of P. vulgata at the | 40 |
|  | six Little Wick Transects |  |
| Figure Twelve | Standard "Beer Bouguer Law" Plot of Absorbance vs. | 19 |
|  | Concentration of oil in Solvent |  |
| Figure Thirtee | en Calculated Regression Lines of Body Burden on | 48 |
|  | Dry Weight for each Transect and for all the data |  |
|  | pooled irrespective of Transect Origin |  |

## List of Tables and Illustrations (continued)

Page
Table I Results for the Statistical Analysis on the Limpet ..... 37
Volume Index Data
Table II Spectrophotometry Data for Patella vulgata ..... 21
Table III Results from the Statistical analysis of the ..... 45
Spectrophotometry data given in Table II for
P. vulgata soft tissue
KEY: To the Symbols used in Tables I. II and TII ..... 38

## INTRODUCTION

After the end of the Second World War, a search for ports capable of accommodating bulk carriers and tanker's of up to 100,000 ton's capacity revealed that Milford Haven was one of the very few suitable sites in Britain. Consequently the remote natural harbour at the south-western extremity of Wales (see figure one) used by fishing vessels, coastal freighter's and small naval craft became Britains largest oil port, a development greatly accelerated by the closing of the Suez Canal in 1956. Milford Haven is situated where the ranges of Arctic and Southern Marine invertebrates overlap and it supports one of the most varied fauna and flora in the British Isles. The great depth of the water channel, so imperative to its functioning as a Modern Oil Port, also means that this variety extends far into the estuary.

A preliminary report by Arnold (1959) and a major account of the estuary by Nelson-Smi.th (1964) provide the foundations of the pre-industrial monitoring of Milford Haven's Marine Biology. Since the early 1960's, when industrial development saw the establishment of three refineries (Esse in 1960, Texaco in 1964 and Gulf in 1968), the monitoring has continued. Paper's on the Physical structure of the estuary and its Marine Biology were prepared by Nelson-Smith (1965, 1967 respectively) and a more detailed study of the Dale peninsular was carried out by Moyse and Nelson-Smith in 1963. Although changes in the intertidal fauna and flora were apparent, these were attributed


Figure One: South West Wales, showing Milford Haven and the surrounding area
to a general climatic deterioration. Furthermore, there was no evidence to suggest that a general impoverishment of the fauna and flora had taken place, and it was suggested that the establishment of an oil port had not caused any major ecological damage to the estuary.

A major re-surveying of all the Milford Haven transects is being carried out at present by the Oil Pollution Research Unit (O.P.R.U.) at Orielton Their results, when compared with the pollution histories of the various areas, should prove extremely interesting and useful in assessing the impact of extension of industrial operations in the estuary since $19 \overline{9} \overline{7}$.

Pollution associated with oil developments can take one of three forms:1. Spillage of oil in the sea. Depending on the type of oil involved, this may lead to deaths of marine organisms, particularly if large percentages of volatile hydrocarbons are present. However, by the time the oil drifts onshore the toxic light fraction's may have evaporated.
2. Spillage of oil directly on intertidal areas. Here stranded oil may kill intertidal animals either by poisoning or smothering. Again the toxic fraction/s of oil disappear rapidly through evaporation and solution, in this way spilt oil will soon loose its toxic properties. Oil is rarely stranded in quantities sufficient to kill intertidal species by smothering, though applications of emulsifier prove far more toxic to, for example Patella vulgata than does the original oil spill (Crapp, 1970).
> 3. Discharge of oil/water mixtures from refinery, or stabilizing tanks, onto intertidal areas.

The first and second categories are examples of 'Acute pollution', the third, of 'Chronic pollution'. The effects of Acute pollution are well documented, (e.g. Dudley, 1968; George, 1961; Nelson-Smith, 1968 (a) and (b), 1970:) but there is much less information on the effects of continuous low level pollution as may be seen around discharge points, where although a very low concentration of oil in water is released, the toxic fractions of the oil mav be continually oresent.

In the case of the above-mentioned refineries, the effluent is derived from three separate sources.

1. Process Effluent - water condensed from the steam injected into the refinery process
2. Fresh Water runoff - from rain falling into the refinery area
3. Ballast water from tankers

Before it is discharged, the effluent passes through skimming pools and separators. Even so, contaminants such as sulphides, copper, cyanides, Phenols \& $c$. and Ammonia are discharged, though below the limits set by the South West Wales lc River Board, and normally amount to a fraction of a mg/litre of the effluent. Oil, the principal contaminant, is limited to $50 \mathrm{mg} /$ litre though normally the actual amount released varies from 20 - $25 \mathrm{mg} /$ litre.

A total of $1 \times 10^{9}$ gallons of effluent may be discharged during a year which may contain up to $20,000 \mathrm{gallons}$ of oil, at a concentration of $20 \mathrm{mg} / \mathrm{litre}$.

The major stimulus for the present investigation came from work carried out by Crapp in the three years prior to 1970. A summary of his findings is given below.

Crapp's work showed that continuous low level pollution, outlined above, was having a significant ecological effect at the Esso discharge point in Little Wick Bay, near the town of Milford Haven (figure two). Crapp visited the bay in 1969, after about ten years of discharging and discovered that Fucus vesiculosus was the dominant intertidal species. If the bay was sufficiently sheltered to be Algal dominated then the main weed should have been Ascophyllum nodosum, however the position the bay occupied within the Haven, as regards exposure to wave action, indicated that it should be a Barnacle/Limpet dominated shore. Indeed photographs taken before industrialisation started in 1960 show this to be the case (see Milford Haven Conservancy Board Booklet 1968).

Crapp investigated why this should be and the details of his experimental work
and findings are discussed fully in later sections. Briffely, Crapp investigated the six transects shown in figure three and recorded the relative abundance of the species present at each. The pattern revealed by the survey indicated that the pre-industrial species distribution of the shore had been disturbed, and the disturbance centred on the outfall adjacent to transect three. It was concluded that the fauna of the shore had been severily depleted and this had allowed invasion by fucoid algae.


Milford Haven


Figure Three: The Little Wick transects in relation to the Outflow Pipe and Esse Jetty

Given that any differences in Topography, Climate and Exposure (Ballantine's scale 1961) between the transects were insufficient to explain the ecological differences noted, Crapp concentrated on unnatural factors for an explanation. (Again a more detailed breakdown of these factors is given later).

The use of emulsifiers to clean Little Wick Bay could have produced the observed distribution of intertidal species but Crapp's enquiries showed that no such cleaning procedure had been used within the few years piat?
sübsequent to his investigation. tren if such a cleaning operation had taken place it is difficult to reconcile such finding's with a general cleaning of the Bay, as the species changes so intimately related to the outflow itself, and Crapp subsequently concentrated on the outflow and its contents.

The use of emulsifiers in the skimming pools and separation plants are a possible cause of the changes observed but Crapp considered this to be unlikely as the usage was reported as highly infrequent (the last known occasion being seven years before Crapps investigation was undertaken). Salinity changes due to the effluent are not great enough to explain the Biological differences recorded, no salinity figures below $27 \%$ o were recorded and the average reading approximated to $30 \%$. It is possible however that reduced salinities may impose an additional stress upon many species.

The possibility therefore remained that it was oil itself, discharged in the effluent, that was causing the changes in species composition observed at transects three (and four to a lesser extent). It seemed unlikely that the other chemical constituents were responsible as the concentrations released were considered low enough to be ineffectual (see Appendix one). For the oil to be implicated it had to be established that continuous oil pollution could have such. a toxic effect at such low concentrations. Laboratory experiments undertaken by Crapp, on Asterias rubens and Carcinas maenas: showed that the extent to which different species might be affected varied considerably. However field studies into some of the dominant intertidal organis/revealed firstly, that first year age class Patella vilgata (taken to be those with a longest shell diameter of less than 5 mm ), were reduced in abundance near the outfall (transect three), and secondly, that Barnacle spats were less abundant at transect three. Knight-Jones (1953b) has shown that Balantas balanoides, B. crenatus and miminius modestus all exhibit a gregarious settling behaviour i.e. the spat is less likely to settle in an area devoid of, or lacking in barnacles of the same species, other factors being equal. Therefore this might be a contributory factor to the low abundances recorded. However Crapp has shown that mortality rates amongst those spats that do settle are higher at transect three than at any other transect in Little Wick Bay. Crapp concluded that some deleterious influence was affecting the young limpets and barnacle spats, settling near the outfall, and the normal refinery effluent was implicated.

The aim of the present investigation, some ten years later, was to compare results of repeated surveys of the six transects, firstly with Crapp's 1970 results and secondly, with Petpiroon's unpublished 1978 results (Taken from his Ph.D. thesis, to which he has kindly allowed me access), and to relate differences or similarities in the findings to the environmental agencies (natural and otherwise) that have operated since that time. I also carried out a quantitative analysis of Petroleum Oil pollutants on the Limpet $\ell . C$. Patella vulgata by infra red spectrophotometry along with a detailed description of the size classes and abundance of the species at each of the six transects; in an attempt to relate this to the Esso refinery effluent discharge in Little Wick Bay.

## MEIHODS

## I. The Shore Survey

This survey, the third of its kind at Little Wick, was undertaken to ascertain if any changes in the intertidal fauna and flora had occurred since Crapp's original work in 1970, and again since Petpiroon's more recent work in 1978, and if so to try to relate this to known environmental agencies in operation in that period.

The Little Wick transects (figure three) are situated as follows, with respect to the Esso refinery discharge point.

Transect 1 - 400 m West of the outfall

| " | 2 | - | 200 m | " | " | 11 | " |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11 | 3 | - | 0 - | " | " | " | " |
| " | 4 | - | 27 m | Fast | 11 | 11 | " |
| 11 | 5 | - | 1.60 m | " | 11 | 11 | " |
| 11 | 6 | - | 360 m | 11 | " | 11 | 1 |

All three survey's were based on station's established at constant vertical interval's along a transect lying roughly at right angles to the line of high or low water. The methods used correspond exactly to those used by Petpiroon and Crapp unless stated otherwise.

Each transect extends from low water of Spring tides to the first few flowering plants at the top of the shore. The zero level, (or Chart Datum), is usually established by the level of low water on the day of the survey, taken from data in the Admiralty Tide Tables. However, in the present survey the uppermost station of each transect was known exactly having been marked by Petpiroon in the previous year, and I decided to start there and work toward's low water mark to ensure that the position's of sempling station's used in the two survey's coincided exactly.

To ensure accurate resurveying, a seaward compass bearing was taken', in this case from the paint mark, though usually this would be from a map reference point, and a tape laid along the bearing line. In accordance with the past survey's the transect was divided into equally spaced vertical interval's of 60 cm from the top marker, which in the Milford Haven estuary ensured at least ten intertidal station's, the average tidal range being 6.3 metres. This was achieved by a "Crosstaff".

The instrument consists of a siting bar fixed at a standard distance up a vertical pole and levelled by a bubble in spirit. The original instrument was modified by Crapp in 1970 to include a mirror, (figure four), so that the previous station couid be sited and the instrument kept level from the same view point. This allowed shore surveying to be carried out by a single worker instead of the pair needed, prior to the modification.

All stations once established were marked off on the bearing tape enabling an approximate shore cross-section to be drawn at a later date (sec figure ten). Notes were kept at each station of the nature of the substratum and any other points of interest.

When considering the distribution of the intertidal fauna and flora, an area of one square metre was investigated as displayed in figure five. This was in exact accordance with Petpiroon's method and any species "present" but outside the sample area were noted as such on the results sheet. The size of Crapp's sample area differed slightly, (for comparison see Moyse and NelsonSmith (1963), who outline the procedure Crapp followed). Organisms in gullies, on the landward side of pinnacles or occupying crevices were scored merely as "present in gullies" unless that was their only habitat (e.g. Littorina neritoides). Similarly organisms in rock pools were ignored, though this habitat proved to be rare at Little Wick.


Figure Four: Showing the present Model of the "Crosstaff" after modification by Crapp in 1970 (Taken from Crapps Ph.D. Thesis, University College of Swansea)


Figure Five: "The Sample area investigated at each station along the Transect Line

The criteria of abundance used, displayed in full in the Appendix, were those developed by Ballantine, Moyse and Nelson-Smith from the original proposals of Crisp and Southward (1958). Crapp (1970) considered that the criteria for "Abundant" were set too low for several groups and thus added the grades "Super Abundant" (S) and "Extremely Abundant" (Ex), but did not modify the original grades except by setting an upper limit to "Abundant". In brief the categories used were Extremely Abundant (7), Super Abundant (6), Abundant (5), Common (4), Frequent (3), Occasional (2) and Rare (1).

Such criteria of abundance were used rather than any absolute number or densities, because after practice they are easier to estimate quickly in the field and because actual numbers per unit area must differ between organisms according to their size. Most recognizable changes involve orders of magnitude and are thus demonstratable by such criteria.

The checklist of species used was that assembled by Petpiroon, slightly modified from Crapp (1970), consisting of 48 locally occurring species of rocky shores. It was considered unnecessary to take Spirorbis specimens further than genus level due to difficulties of identification and rarity of occurrence. Likewise Limpets were left at genus level as Crapp's study had $<$ e shown Patella aspera and Podepressa to be rare at Iittle Wick and identification involved the removal of the specimens which was considered undue disturbance of the habitat for the extra information gained.

For identification purposes Collins Pocket Guide to the Sea Shore (Barrett and Yoúnge, 1973) proved sufficient in most cases with additional notes from $Q$ O.P.R.U. and Petpiroon (pers. comms.) where necessary.

Previous estimates of the exposure grade of the shore using Ballantine's (1961) scale were accepted as accurate. However as an exercise in familiarization with the species involved, transect five was surveyed to ascertain if the exposure score thus obtained would agree with the literature. The technique involved,
for every organism listed, the circling of every grade whose criteria of abundance corresponded with the maximum found on the shore (see Appendix). The number of 'rings' for each exposure grade were then summed and the exposure grade with the most rings was classed as the grade for that shore or part of the shore. Checking every listed species is important in such a survey as the technique relies on the presence, or absence, of certain indicator species common to the eight grades of exposure observed.

Emphasis has been placed in my study on the exact location of each transect and its subsequent stations. This should enable future survey's to relocate the exact sample points studied previously, and enable comparison's of changes in species composition, with time (and any changes in environmental
 The results of the shore survey are given in full in the Appendix.
II. Studies of Patella vulgata at the Six Little Wick Transects

During the period in which survey's of the six Little Wick transects were carried out, it was noticed that Limpets varied considerably in size from transect to transect. Transect three, and to a certain extent transect four, appeared to support Limpet populations of greater average size (estimated visually) than did transects one, two, five and six.

I therefore decided to investigate firstly, whether the differences in average Limpet size were statistically significant, and secondly, whether any $\hat{\ell}$. differences in body burdens and/or concentrations of Hydrocarbon could be detected in P. vulgata.

Patella vulgata was considered a suitable subject for investigation for two reasons:-

1. It was recorded in significant numbers on all six transects
2. Limpets as a group vary in size under different environmental
regimes (Lewis and Bowman, 1975). This being the case, further insight into possible effects of the Esso refinery effluent may be gained from a study of the genus.
(i) Investigation of the sizes of Patella vulgata in Iittle Wick Bay

From the initial shore survey it was discovered that P. vulgata reached its greatest abundance around station four of each transect studied. It was decided that this would be an appropriate station to adopt for the study of the size variation outlined above.

Sampling proceeded by counting every Limpet specimen within two 50 cm quadrats either side of the bearing tape. In this way half of a square metre was sampled for each transect unless this did not provide adequate numbers of results for subsequent statistical analysis. If this was the case subsequently larger areas were sampled as appropriate. The results are given in figure eleven.

Measurements of the longest diameter, shortest diameter and height of the shell were taken and a shell volume index calculated (see Appendix). For the calculation the Limpet shell was assumed to approximate to a cone. (ii) Quantitative analysis of Petroleum Oil Pollutants by Infra red

Spectrophotometry
The method employed for the determination of Hydrocarbon body burden in Patella vulgata is a relatively new approach to the proklem and one which has not been previously used on this particular genus as far as is known. For this reason the methods are outlined in somewhat greater detail than would normally be appropriate.

The use of Spectroscopic, as opposed to Gravimetric methods in the quantitative analysis of water dispersed oils, is advantageous for a number of reasons.

Both methods employ solvent extraction to isolate and concentrate water
dispersed oils for quantitative measurement. Solvent evaporation, or "stripping", which precedes weighing in Gravimetric methods has the $\operatorname{l}$. drawback that some of the volatile petroleum fractions are lost (Gruenfeld, 1975). Also, questionable sensitivity and accuracy are achieved by weighing minute oil residues in comparitively large ( 125 ml ) distillation.flasks in another gravimetric method (American Public Health Association 1971). Harra and Somersalo (1958) conclude that the Spectroscopic methods are far more sensitive and accurate.

## (a) Sampling and Sample Preparation

Sampling was carried out on the 5th June, 1979 at Low Water Mark (figure six), and transects were relocated as described in the previous sections.
 the tape and its base just above the water line as shown in figure seven. Conditions were calm and the tide was on the ebb side of 'on the turn' i.e. no allowances had to be made for water movement in between the sampling stations.

Sampling proceeded in the top left, or right, division of the quadrat, and from the top left corner of that division, as indicated in figure seven, until three Limpets from three sample points ( $a, b$ and $\underset{=}{c}$ ) had been obtained $\in \mathbb{E}, ~ u . i$ at all six of the transects. Thus any subjective error due to variation in specimen size was reduced to a minimum. The samples were collected in labelled plastic bags and transported to the laboratory where they were removed from their shells and washed in clean sea water. This precaution ensured that no exogenous source of Petrogenic Hydrocarbon (namely oil residue $\ell_{\text {c }}$ on the body or shell of the Limpet, or on weed attached to the shell), was included in the analysis. The samples were then air dried for up to 78 hours allowing the largest of the specimens to dessicate completely. Once fully dried the specimens were weighed and wrapped individually in filter paper for extraction.
(b) Extraction and Analysis
(i) Apparatus

Extractions were performed using standard Soxhleth apparatus, and the Hydrocarbons determined quantitatively with a Pye Unicam SP 1050 Infrared Spectrophotometer. Solution absorbances were all measured in matched 10 mm quartz cells.
(ii) Procedure

Firstly a "Beer-Bouguer Law" plot had to be prepared. Quantitative determination by a"single point analysis" (Gruenfeld, 1975) requires a linear plot that passes through the origin. To obtain such a plot, five standard solutions of accurately known concentrations of oil in Carbon Tetrachloride were prepared. The oil solution concentrations were adjusted to yield absorbances that were within the ordinate scale range of the Infrared Chart Paper.

Zero ordinate scale expansion and 10 min quartz cells required 5 solutions of concentrations ranging between 0.5 and $40 \mathrm{mg} / 100 \mathrm{ml} \mathrm{CCl}_{4}$ and these were prepared accordingly. The solutions were then put through the spectrophotometer (with a "blank" of $\mathrm{CCl}_{4}$ from the same bottle) and the readings converted into absorbance values. These absorbance values were plotted against the known concentrations of oil in solvent and the standard plot prepared (see figure twelve). The absorbance band maxima of the oil mixture used was $2750 / \mathrm{cm}$.

The calculation of absorbance values proceeds as follows: A typical trace from the spectrophotometer is shown in the figure eight.

A baseline $(p ; q)$ is drawn, then the values $A$ and $B$ are read from the infra red paper. The values $A$ and $B$ are then summed to give the transmission value ( $1 \%$ ) which is substituted into the formula given below to obtain the absorbance value.

$$
\text { Absorbance }=\log _{10} \frac{(100)}{(1 \%)}
$$



Figure Six: Showing sample Points from the six Little Wick Transects) (H.W.M. $=$ High water mark)


Figure Seven: Detail of a Sample area for any given transect (L.W.M. = Low water mark)


Figure Eight: An idealised trace from a Spectrophotometer readout


Figure Twelve: A standard "Beer Bouguer Law" Plot (Solution V was contaminated and was not plotted)

Once the transect samples have been processed and absorbance values obtained either a body burden may be read off the standard plot or, more accurately, an arbitrary value is read from the plot and this plus the sample absorbance reading are substituted into the following formulae used for single point analysis.
$C x=C s \quad \cdot \frac{A x}{A s}$
Where $C x=$ the unknown oil concentration of the sample extract used for infra red measurement ( $=$ Limpet body burden)

Ax and As $=$ The absorbances of the sample extract and standard solution respectively.

Cs $=$ The Standard solution concentration used tor $1 . \mathrm{K}$. measurement The results are given in Table II.

A single point analysis as opposed to a full Infra red scan offers considerable time saving which is invaluable in such work requiring over fifty samples to be processed. Before sample absorbances could be put through the above procedure each sample was extracted in 200 ml of Carbon Tetrachloride (B.P. $72^{\circ} \mathrm{C}$ ) for six hours. The resultant solutions were allowed to come off the boil before being decanted into flasks and put through the spectrophotometer. Once again a 'blank' of $\mathrm{CCl}_{4}$ was used taken from the same bottle as that used for the extractions

Table II Spectrophotometry Data for Patella vulgata

| Sample <br> Number | 1\% | Ax | Cx | Dry <br> Weight | [] |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1A(i) | 54 | 0.268 | 18.48 | 0.494 | 37.42 |
| 1A(ii) | 65 | 0.187 | 12.89 | 0.380 | 33.92 |
| IA(iii) | 86 | 0.066 | 4.55 | 0.278 | 16.37 |
| 18(i) | 43 | 0.366 | 25.24 | 1.389 | 18.17 |
| 18(ii) | 89 | 0.051 | 3.52 | 0.148 | 23.78 |
| 18(iii) | 70 | 0.155 | 10.68 | 0.530 | 20.15 |
| 1c(i) | 51 | 0.292 | 20.14 | 1.057 | 19.05 |
| 1C(ii) | 77 | 0.114 | 7.86 | 0.299 | 26.29 |
| 10(iii) | 76 | 0.119 | 8.20 | 0.454 | 18.06 |
| 2A(i) | 50 | 0.301 | 20.75 | 0.624 | 33.25 |
| 2A(ii) | 58 | 0.236 | 16.28 | 0.546 | 29.82 |
| 2A(iii) | 63 | 0.201 | 13.86 | 0.413 | 33.56 |
| 2B(i) | 31 | 0.509 | 35.10 | 0.758 | 46.31 |
| 2 B (ii) | 67 | 0.174 | 12.0 | 0.450 | 26.67 |
| 2B(iii) | 77 | 0.114 | 7.86 | 0.193 | 40.73 |
| 20(i) | 55 | 0.259 | 17.86 | 0.627 | 28.48 |
| 2c(ii) | 56 | 0.252 | 17.38 | 0.446 | 38.97 |
| 26(iii) | 61 | 0.215 | 14.83 | 0.482 | 30.77 |

Mable II (continued)

| Sample <br> Number | T\% | Ax | Cx | Dry <br> Weight | [] |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 3A(i) | 26 | 0.585 | 40.34 | 1.059 | 38.09 |
| 3A(ii) | 29 | 0.409 | 28.21 | 1.441 | 19.58 |
| 3A(iii) | 32 | 0.495 | 34.14 | 1.285 | 26.57 |
| 3B(i) | 54 | 0.268 | 18.48 | 1.029 | 17.96 |
| 3B(ii) | 29 | 0.538 | 37.10 | 2.697 | 13.76 |
| 3i(iii) | 34 | 0.468 | 32.28 | 1.448 | 22.29 |
| 36(i) | 43 | 0.366 | 25.24 | 1.314 | 19.21 |
| 30(ii) | 37 | 0.432 | 29.79 | 1.720 | 17.32 |
| 3C(iii) | 34 | 0.468 | 32.28 | 1.171 | 27.57 |
| 4.A(i) | 44 | 0.356 | 24.55 | 0.825 | 29.76 |
| 4A(ii) | 65 | 0.187 | 12.89 | 0.445 | 28.97 |
| 4A(iii) | 65 | 0.187 | 12.89 | 1.209 | 10.66 |
| 4B(i) | 52 | 0.284 | 19.59 | 1.019 | 19.22 |
| 48(ii) | 44 | 0.356 | 24.55 | 1.165 | 21.07 |
| 4B(iii) | 69 | 0.161 | 11.10 | 0.950 | 11.68 |
| 4C(i) | 43 | 0.366 | 25.24 | 1.078 | 23.41 |
| 4C(ii) | 67 | 0.174 | 12.0 | 0.465 | 25.81 |
| 4C(iii) | 29 | 0.538 | 37.10 | 1.880 | 19.73 |
| 5A(i) | 62 | 0.208 | 14.34 | 0.496 | 28.91 |
| 5A(ii) | 82 | 0.086 | 5.93 | 0.200 | 29.65 |
| 5A(iii) | 75 | 0.125 | 8.62 | 0.349 | 24.01 |

Table II (Continued)

| Sample <br> Number | T\% | Ax | Cx | $\begin{gathered} \text { Dry } \\ \text { Weight } \end{gathered}$ | [] |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5B(i) | 66 | 0.180 | 12.41 | 0.823 | 15.08 |
| 5B(ii) | 81 | 0.092 | 6.34 | 0.325 | 19.51 |
| 5B(iii) | 33 | 0.481 | 33.17 | 1.756 | 18.89 |
| 5c(i) | 69 | 0.161 | 11.10 | 0.661 | 16.79 |
| 56(ii) | 48 | 0.319 | 22.0 | 0.672 | 32.74 |
| 50(iii) | 71 | 0.149 | 10.28 | 0.4206 | 24.02 |
| 6A(i) | 61 | 0.215 | 14.83 | 0.454 | 31.67 |
| 6A(ii) | 72 | 0.143 | 9.86 | 0.329 | 29.97 |
| 6A(iii) | 60 | 0.222 | 15.31 | 0.284 | 53.91 |
| 6B(i) | 73 | 0.137 | 9.45 | 0.337 | 28.04 |
| 6B(ii) | 68 | 0.167 | 11.15 | 0.409 | 27.26 |
| 6B(iii) | 68 | 0.167 | 11.15 | 0.290 | 38.45 |
| 6c(i) | 58 | 0.236 | 16.28 | 1.198 | 13.59 |
| 6c(ii) | 62 | 0.208 | 14.34 | 0.438 | 32.74 |
| 6c(iii) | 70 | 0.155 | 10.69 | 0.184 | 58.09 |

## DISCUSSION

## 1. The Shore Survey

(i) The distribution of Littoral Animals and Plants in 1979 in relation to the refinery effluent discharge point (see appendix)

## Flowering Plants and Lichens

The presence or absence of flowering plants has been recorded for complethess, but they are considered to be of little relevance to the present investigation. Similarly some of the more characteristic lichens of the "supra littoral" zane,. (Lewis, 1964) have been noted. Any distributional differences between the transects $6 f$ which there seem to be none of distinction) are considered to be for reasons other than those attributable
 as "Verrucaria mucosa", show a reduction in relative abundance and a reduction in the width of the zone of occurreneeat transects three, four and, to a lesser degree, two, (the three transects nearest to the outfall pipe.) The Brown Algae (Phaeophyceae)

A pronounced increase in the relative abundance of Pelvetia canaliculata was observed at transects three and four. Po canaliculata was absent at transects one and two, and present in reduced amounts at transects five and six (figure nine).

Both Fucus spiralis and Ascophyllum nodosum were absent from transects one, two, five and six, which contrasts with the situation on transects three and four (figure nine). A. nodosum shows an increased relative abundance and zone of occurrence at transect three whereas F. spiralis shows this at transect four.

Laminaria digitata was present at all six transects. The reduction in relative abundance observed at transects three and four may be attributed to the instability of the substration which is mainly sand and boulders at the lowest stations investigated.


0





Figure Nine:
Distributions of six common species at the six Little Wick transects. (The abundance scale follows the format outlined in the methods section)


Fucus serratus showed a maximum relative abundance at transect three and a greatly increased zone of occurrence, the same was true for transect four but to a lesser degree. At transect five the relative abundance was similar to transect four but with a slightly narrower zone of occurrence whilst at transects one, two, and six the relative abundances and zones of occurrence were markedly reduced. F. vesiculosus exhibited the same trend but was less abundant than F. serratus and absent at transects one and six.

## The Green Algae (Chlorophyceae)

Species of Ulva were rarely found at any of the Little Hick transects, outside the rock pools, though a small increase in relative abundance was noted at transect three. Enteromorpha species were found at all transects and the zone of occurrence widened at transects three and four. However the relative abundances were similar throughouta

## The Red Algae (Rhodophyceae)

Gigartina stellata showed a slight increase in relative abundance at transect three though the distribution among transects was variable. The reduction of Gigartina stellata and the Rhodophyceae as a whole at transect four may be explained by the somewhat unstable nature of the substratum between stations three and five consisting of sand and boulder's. No other general trends were apparent with the exception of a reduction in the relative abundance of Porphyra at transect three. Barnacles (Cirripedia)

In contrast to the situation at transects one, two, five and six, Chthamalus stellatus was noticeably reduced in relative abundance, and had a much narrower zone of occurrence, at transect four, and was unrecorded at transect three (figure nine). Eminius modestus however showed no reduction in its zone of occurrence at any of the transects, and only a slight reduction in abundance at transect three.

Balanus balanoides showed a marked reduction in relative abundance and a restricted zone of occurrence at transects one, two and three. The situation was alleviated slightly at transects four and five, and a maximum relative abundance, and zone of occurrence were seen at transect six.

Balanus crenatus occurred only in small numbers at any of the transects (excepting transect six) and is noted as rarely extending above the infra littoral fringe (Lewis, 1964).

## Winkles (Littorinidae)

All species of winkle recorded at Little Wick showed marked reductions in both abundance and zone of occurrence at transects three and four, Littorina neglecta being totally absent there (figure nine).

## 

The Topshells reflected the same distributional changes as the Littorinid's, but more drastically. All species of Topshells were completely absent at transects three and four. Maximum abundances for Gibbula umbilicalis and Monodonta lineata were recorded at transects six though the zone of occurrence remained fairly constant.

## Limpets (Patellidae)

Limpets were separated only to genus level (for reasons' given above). Both reduced relative abundance and zone of occurrence were observed at transects three and four (figure nine). The other four transects supported more abundant populations. (The next section deals with Patella vulgata in more detail). The remaining Fauna

Mytilus edulis was recorded in small numbers at transect six only. Nucella lapillus the dog whelk was recorded at transects one and six only, again in small numbers.

The Serpulid Pomatoceros triqueter was rarely recorded as it is confined to the infra littoral fringe. The coelenterate Actinia equina was absent at transect treee
though present in varying abundances at all other transects.

## (ii) Possible Explanations of the observed Species Distributions

In general the central transects at Little Wick (transects three and four), differ from the other four transects firstly, by supporting a greater number of species of Brown Algae and a greater percentage cover of the majority of those species, and secondly, by supporting a depleted number of faunal species and a lesser abundance of the majority of those species.

## The Fauna

With the exception of Balnus balanoides the species of barnacle found at Little Wick reach their lowest abundance at the central transects. (B. balanoides is actually less abundant at transect one and this remains essentially unexplained). Trap (1970) implicated the refinery effluent as the main cause of such reduced abundances and noted reduced numbers of barnacle spats plus an increased spat mortality rate near the outflow. Knight-Jones (1953b) has shown that $\mathrm{B}_{0}$ balanoides, $\mathrm{B}_{\text {e conatus }}$ and EMminius modestus spats show a gregarious settling behaviour. Thus less spats settle in areas of low barnacle density, all other factors being equal. The reaction was show to be species specific. More recent work by Dick's (1975b) in Little Wick Bay has shown that fewer B. balanoides spats settle near the outfall (none having settled on the end of the pipe), and that the numbers increase with distance from the outflow. Ib establish why this might be, Dick's carried out laboratory tests on the nauplii of B. balanoides and concluded that the Little Wick effluent had a dual influence on barnacle larvae, firstly because of its reduced salinity, and secondly because of "other effluent constituents which were not measured". The barnacle nauplii are positively phototactic and thus swim to the surface of the sea; the effluent also forms a thin layer on the surface of the water. When the nauplii encounter the effluent, an immediate cessation of swimming occurs and the nauplii drop below the effluent layer. This observation explains
the observed pattern of spat settlement as follows - As the surface film of effluent moves up and down the shore with the tide,so spat settlement is prevented in the area the effluent covers. (NB Although it is the cyprid stage that settles and not the planktonic nauplii larvae, it is assumed, not unreasonably, that it reacts to the effluent in the same way.) However, there are long periods at high tide when the effluent has no effect on the middle shore during which time some cyprids will have a chance to settle. Hence Dick's concluded that settlement could only be seriously influenced where the effluent occurred continuously, very close to the outflow pipe, with a lessening effect as distance from the outfall pipe increases, produced by dispersion of the effluent. It seems reasonable to conclude from Dick's work that the effluent effects spat settlement, but that once settled this barnacle may continue growth and development normally.

The settlement of Chthamaltus stellatus may also be affected in the manner suggested above. However, the complete absence of C. stellatus at transect three suggests the effect might be even more acute. As well as the nauplii and cyprid showing aversive reactions to the effluent, there might be mortality among both the young stafes and adults.

The results for Blminius modestus require another explanation. Compared to the other barnacle species present at Little Wick, E. modestus is comparatively abundant at transects three and four, and it is possible that Elminius' settling stage is less sensitive to the effluent. It has also been established that Himinius grows well underneath a covering of Fucaceae (NelsonSmith, 1967), therefore once established it may be in a position to outcompete other barnacle species less well adapted to growth under a fucoid canopy, such as B. balanoides and Chthamalus stellatus (Lewis, 1964).

In short, the observed reduction in barnacle densities near the outflow pipe (transects three and four) may be caused directly by the effluent itself,
and indirectly by the cover of fucoid algae which have invaded the area.
The winkle population at Little Wick showed an extremely marked change at transects three and four. As early as 1968 L. Iittorea and L. Iittoralis were recorded as being very sensitive to oil pollution (Nelson-Smith). Parsons, (1972a) noted that L. saxatilis ceased activity and retracted into its shell when exposed to effluent and Baker (1975b) has shown that $50 \%$ of a population of I.saxatilis in a rock pool, crawled up and out of the water when effluent was added.

A mark and recapture experiment was carried out by Baker (1975b) in which she released one hundred L. littorea near the Kent refinery outlet. Nn uinkloo wara mononturat noar the offluent nine: and those ohserved fust after release were in a "sick flaccid" condition and presumed to have been subsequently washed away. This confirms Parson's observations, as retraction into the shell was often followed by the winkle being washed out of the area. Recovery did not ensue uniil the water was free from effluent or the animal was carried out of the contaminated area (Parsons 1972a; Baker, 1975b).

Nelson-Smith (1968) observed that when L. neritoides lives in dead barnacle cases, the oil pollutants tended to wash harmlessly past, thus protecting an otherwise very vulnerable species. This observation was unlikely to be of importance at Little Wick as no winkles were recorded in dead barnacle shells in the central transects.

Although L. Iittorea is rare at Little Wick Bay (Nelson-Smith, 1967) it is interesting to note that the free swimming larval stage avoids surface waters of low salinity (Brattegard, 1966). Referring to the observations for B. $_{\text {. }}$ balanoides nauplii larvae, reviewed above, this may explain the reduced relative abundance of $I_{\text {. }}$ littorea at transects three and four compared to transects five, six and one.

The effects of refinery effluent on the topshells Gibbula umbilicalis
and Monodonta lineata have not been studied. However as the refinery effluent has a direct effect on Barnacles and mollusc species closely related to topshells, namely Littorinids, it is likely that the complete absence of either of the $\ell . c$ above two topshells is due to the effects of the refinery effluent. NelsonSmith (1968) cites M. lineata as another sensitive species to oil pollution. Crapp (1970) has found that both topshell's and winkles retract into their shells when under the influence of pollutants and thus are washed into the sublittoral zone, eventually crawling back onto the littoral zone in a nonpolluted area to become reestablished.

Lewis (1964) notes that dense shading, encountered for example under a dense cover of fucoid algae, favours Gibbula umbilicalis, Monodonta lineata
 central transects (all being absent from transect three). Thus, presumably, sensitivity to the effluent negates any beneficial effects of the fucoid cover.

Mytilus edulis, Nucella lapillus and Pomatoreros triqueter are so rare at Little Wick that any attempts to relate their distribution to the effluent would be dubious. N. lapillus might be expected to be reduced or absent from the central transects due to reduced densities of its prey species (barnacles). regardless of any direct effect of the effluent.

The possible explanations for the reduced Limpet densities are dealt with $\ell . e$ fully in the following section.

Another trend observed from the Little Wick transects was the increased abundance of Nucella lapillus, Monodonta lineata and Gibbula umbilicalis at the eatern transects five and six. As noted in Crapp's study these transects are marginally more sheltered than transects one and two. The above mollusc's are reported to be better adapted to sheltered areas within Milford Haven (Nelson-Snith 1967) and this may explain the observed distribution.

The Flora
With the exception of Luminaria digitate (whose main abundance was outside the sampling area), all of the Phaeophyceae increased in abundance and showed an increased zone of occurrence at the central transects. Algae are reported to be unusually resistant to effluent because they are protected by a mucilage covering. The abundance of Brown Algae within the immediate area of the refinery discharge shows them to be resistant (Baker, 1975). However resistance to effluent toxity is not enough on its own to explain the increases in abundance observed. The main factor responsible for the observed increase in brown algae must be reduced grazing pressure from Limpets, (Trap, 1970; Baker, 1975). Reduced
 the algae have reached maturity they are extremely resistant to grazing damage, except by very large limpets.

Enteromorpha shows an increased abundance and zone of occurrence at the central transects. An abundant growth of these green seaweeds was reported by Baker (1975a), followed by a growth of fucoid algae following Limpet detachment, shortly after an oil spill in the Milford Haven estuary. This particular green algae is noted for its great resistance to effluent discharge (Baker, 1975), and Nelson-Smith (1967) has reported Enteromorpha to be abundant at all levels where fresh water drains across the shore. As exposure to effluent also coincides with exposure to waters of reduced salinity, the ability of Enteromorpha to tolerate salinity changes may in part explain its resistance to effluent discharge.

Van Gelder-Ottway (1975) has looked at the effect of a floating oil film: on algal Photosynthesis and concludes that for frtermorpha neither gas exchange $l_{c}$ temperature or light reduction effects are significant in the marine environment. (Some effects were observed in the Pockpool habitat though).

The increase in relative abundance of Gigartina stellata at transect three is probably due to the dense fucoid cover which encourages many red
algae to extend their range further upshore．The fucoid cover prevents dessication，a condition to which the Rhodophyceae are particularly susceptable．

Porphyra actually decreases in abundance at transect three．Since Porphyra is a weed of the more exposed areas of the Pembrokeshire coast （Bvans，1947）it is possible that the fucoid cover offers too sheltered an environment for the successful establishment of this algae．Such observations are supported by the fact that Porphyra attains its greatest relative abundance at transect one，the most exposed of the six at Little Wick，having an exposure grade of four（Ballentine，1961）．
（iii）Comparisons of the present Survey with the Survey＇s of Crapp（1970） and Petpiroon（1978）

Not surprisingly the differences between Petpiroon＇s results and my own， （taken one year apart and at the same time of year），are negligible．

Before comparing the differences between Crapps results and my own（taken nearly ten years apart and at the same time of year），it will be useful to summarise Crapp＇s conclusions（full results of both Surveys are given in the Appendix）．

1．There is a change in exposure from South Hook Point（Grade three， Balldntine，1961）to Gelliswick（Grade six）．The Little Wick transects vary／a between grades four and five．The Bay itself（transects three and four）is probably more sheltered（Grade five）as is transect six，while transect＇s one， two and five may be assigned to Grade four．Changes in exposure are not sufficient to explain the variations found in flora and fauna．

2．Several species of gastropod molluscs were absent or reduced in numbers on the central transects．This could not be explained in terms of natural environmental factors．Littorina saxatilis and Littorina littoralis appear to be unaffected this way．

3．Limpets（Patella vulgata）and barnacles（Chthamalus stellatus）Balanus
balanoides and ⿴⿱冂一⿱一一厶儿立inus modestus）were considerably reduced in numbers on transects
three and four.
4. The Seaweed's Fucus serratus, F. vesiculosus and F. spiralis were particularly abundant on the central transects. Pelvetia canaliculata and supralittoral lichens did not appear to be affected.

Trap's hydrographic studies showed that surface currents away from the outfall are slow, except on the early stages of the ebb and the flood. It is possible that the siting of the outflow leads to some retention of effluent in the Bay.

Crapp's survey and my own are in close agreement in general but there are one or two differences apparent.

Patella vulgata has been further reduced at transects three and four since 1970, as has Chthamalus siellatus at transect three. If the effluent is reducing recruitment, which is strongly indicated by the evidence given above, and that in the next section, then it is not surprising that numbets of Patella and Chthamalus are falling even if the mortality rate has remained constant, though this is only likely to apply if both species are very long lived and longevity is discussed later.

Since the earlier survey, Balanus balanoides has been reduced in relative abundance at all six transects, the greatest effect being shown at transects one, two and three. The relative abundance of Chthamalus stellatus is also reduced at transect one. $\frac{\text { Minus modestus }}{\Omega}$ has an extended breeding season, a higher growth rate, a high dessication tolerance, and adaptability to variations in temperature and salinity (wait 1968). It is likely that <compat>ᄑminus is outcompeting B. balanoides for space at all transects, B. balanoides is better adapted to more sheltered conditions so it is possible that any competitive effects might be greater at more exposed areas thus explaining the above distributional changes. However no subsequent increase in GHinius has occurred since 1970 so this rather weakens the hypothesis.

The ability of minimus modestus to tolerate variations in temperature
and salinity might explain why it $i s$ less affected by the effluent than the other barnacles species. The observed mainiusmortalities probably result only from the toxic chemicals in the effluent (Crapp's data implicates the oil fraction of the discharge). B. balanoides was affected both by the toxic chemicals in, and the reduced salinity of the effluent (Dick 1975b).

Contrary to Crapp's findings, both Littorina saxatilus and L. Iittoralis were reduced at transects three and four in the present survey for reasons which have been suggested above I. littoralis numbers have also decreased since 1970 at transects five and six, this correlates well with a reduction of Fucus vesiculosus and Fe spiralis at the same transects, (Le littoralis is a grazer on fucoid algae). The Fucus decrease, however, is less easy to explain as it does not correlate with increases of Limpets or other grazing molluscs. $\ell C$. Possibly at the time of the earlier survey the Fucoids were abundant following $\ell . C$. a localised oil spill (Baker, 1975). Since then, if young Iimpets have l.t. reestablished, they would check the growth of new Fucus plants. The overall abundance of the algae would decrease as the older plants died.

The final difference between the two survey's is that Crapp maintained Pelvetia canaliculata was essentially unaffected by the effluent. My observations indicate that it has increased its relative abundance at transects three and four, and decreased its relative abundance at transects one, two, five and six, since the 1970 survey. The reasons for these changes may be similar to those for the other brown Algae.

## Conclusions from the Shore Survey

Crapp implicated the continuous input of low levels of crude oil into Little Wick Bay, as being the major factor influencing the observed biological differences apparent between the transects investigated. In doing so he rejected the possibility that ephemeral discharges of emulsifier in the effluent were responsible.

The fact that the three surveys of the Little Wick transects give essentially the same results suggests two important conclusions.

1. The mean Effluent quality has not changed significantly since 1970 , except $\ell . c$ for the fact that oil content has been reduced from a mean of 25 mg /litre to one of 15 mg /litre (Iso 1979, pars. comm.). The survey's show that the situation at Little Wick is essentially an Equilibrium one. Thus a relatively lac. constant effluent is producing a relatively content Biological effect and this supports Craps early conclusions that continuous low level pollutants are the main vectors in bringing about the observed changes.

The implication is that the oil is the actively toxic constituent of the effluent (Trap, 1970). The effect of reduced salinity is shown to be important for some species.
2. The area affected by iso's discharge into little Wick Bay has not increased over a period of ten years since industrial operations started, and the effects are localised to the immediate vicinity of the outflow pipe.
2. Investigations concerning the limpet Patella vulgata
(i) Comparisons of the Size-frequency distributions of P. vulgate at the
six Little Wick Transects
There was sufficient doubt as to whether the data in this section met the assumptions of analysis of variance, to make the use of non-parametric methods necessary. Here the null hypothesis is not concerned with specific parameters (such as the mean in analysis of variance) but only with the distribution of the variates.

A Kruskal-Wallis test showed that the limpet samples differed in 'location' significantly (Sokal and Rohlf, 1969). Once this had been established, fifteen Wilcoxon two-sample tests, comparing each transect with every other transect, were carried out. These tests established whether, for instance transect one had the same distribution of Limpet sizes (volumes) as transect three or not, and the results are as given in Table I.

If we look firstly at the results differing at the $P=0.01$ level of significance, the Limpet distributions at both central transects were

Table I Results for the Statistical analysis on the Limpet Volume index data (for the raw data see Appendix)


Key to the Symbols used in Tables $I_{2}$ II, III

| Symbol | Meaning |
| :---: | :---: |
| H | Kruska-Wallis H statistic |
| H/D | Corrected value for tied ranks |
| P | Probability Level |
| $t_{s}$ | Sample statistic of $t$ distribution |
| 5 | Significant |
| N.S. | Not Significant |
| T\% | Transmission |
| $\dot{A x}$ | Ädsordance vaiue |
| cx | Hydrocarbon Body Burden (200 ml. $\mathrm{CCI}_{4}$ |
| [] | Hydrocarbon Concentration $C x /$ Dry Wt. |
| $\mathrm{F}_{\mathrm{x}, \mathrm{y}}$ | Fishers ratio with $x$ indexing Transects df |
|  | Fishers ratio with $y$ indexing error $\mathrm{ff}^{\text {d }}$ |
| T ${ }_{\text {n }}$ | Transect number (1 $\longrightarrow$ 6) |
| $\mu$ | Mean |
| $\tau_{i}$ | Treatment ie. Transect origin |
| eij | Error |
| df | Degrees of Freedom |




0


Figure Ten : Approximate Shore Cross-sections for the six Little Wick Transects. (The height of station one above chart datum is indicated by the arrows)


 ※


T.4. M.V. $=3 \cdot 424 \mathrm{~cm}^{3}$ $D=124 / \mathrm{m}^{2}$

SHELL VOLUME ( $\mathrm{cm}^{3}$ )

Figure Heven: Size-Frequency distributions for P. vulgata at the six Little Wick Transects. (M.V. = Median volume, D = Density)
significantly different from those of any other transect.
Crapp (1970) and Baker (1975) observed:-

1. Limpet densities per $\mathrm{m}^{2}$ measured at mid tide level show that densities are lowest near the effluent in Little Wick Bay, and that the largest limpets occur where the density is lowest.
2. The youngest limpet classes were missing from these areas characterised by low density and large size. Ovary weights indicated that the large limpets near the effluent were healthy.

These findings correspond exactly with the results of the present survey and once again show the situation to be essentially unchanged over a ten year period. Of relevance here is Lewis and Downan's (1975) obsorvation that relative densities of ${ }^{\text {P. vulgata }}$ from different habitats remain fairly steady as long as the biological condition have remained similar.

The evidence from the shore surveys strongly suggests that any biological differences observed at the Central Little Wick transects are as a result of effluent discharge into that area. Although no direct evidence is available from experiments with Patella vulgata and refinery effluent, once again the inference is that variations in Iimpet size/frequency distribution are due to the effluent. In support of this Dick's (1975b) has shown P. vulgata to be very sensitive to both crude oil and dispersants.

Baker (1975b) transplanted limpets (in the size range $20-40 \mathrm{~mm}$ longest diameter) onto the end of the Little Wick effluent pipe and to control areas, to find if there was an effect of effluent on adult animals. She concluded that there was not such effect. It therefore seems likely that as for B. balanoides (Dick, 1975b) the effluent affects limpet settlement or young stages and the absence of young age classes on the central transects supports this hypothesis.

The release of Emulsifiers through the effuent output is rare (Crapp, 1970). (. However the occasional release might kill a few of the adult population as

Dick's (1975b) has shown that P. vulgata exposed to emulsifier will drop off the rock and are then more easily subject to predation at the next low tide. Experiments to confirm this, (Baker 1975b) showed that when individuals were marked and oiled they dropped off the rock's and were no longer present in the area on the second low tide after the oiling. Dick's (1973) has also shown that $P_{\text {. vulgata }}$ is more susceptable to pollutants at certain times of day. The limpet exhibits a diurnal pattern of activity, being most active at midnight and least active during the day 'superimposed on this is the tidal rythm). Activity, feeding or otherwise, involves the limpet leaving its home scar and Dick's work showed that $60-64 \%$ of a population of limpets detached from a rock when oiled while feeding, whereas only $15-24 \%$ detached from home scars when oiled, then it is unlikely that the normal effluent will affect the adult limpet populations in this way but the effluent quality varies and a pulse of high oil content or a release of Emulsifief, however, rare may be lc. $h^{r}$ a contributory factor.

There are a number of reasons why limpets might attain a larger size near the effluent outflow. Lewis and Bowman (1975) have found an inverse relationship between density and mean size in P, vulgata. Reduced intraspecific competition and an abundant food supply found at transects three and four might well contribute to a faster growth rate and hence to a larger size.

Alternatively the limpets at the central transects may be older than those at transects one, two, five and siz. Very little is known about the age attained by shore organisms but sixteen years is considered feasible (Lewis and Bowman, 1975). Baker (1975) suggested that the limpets at the central transects became established before industrialisation in 1960 and had attained their large size from the availability of an abundant food supply for many years. This explanation is unlikely for all but a few individuals as nineteen years have passed since industrialisation began, and the survival of a large number of individuals for that time span must be improbable. It is more
likely that some limpets having settled further away from the effluent piper; will subsequently be able to move into the area and establish themselves near the outflow. Once established, the abundant food supply and reduced intraspecific competition will allow a large size to be attained.

Interspecific interactions may also contribute to the large size obtained by Pe vulgata near the outflow. Lewis and Bowman (1975) have found the highest growth rate and maximum length obtained when barnacle density is lowest (as is the case here). This is also the case near a dense Fucus canopy as suggested above and also under a Fucus canopy due to increased shelter and reduced barnade density. Lewis et al. have also correlated a high survival rate and elongated life span amongst Pe vuigata where annual recruitment is low, sum Baker's point about greater longevity at transects three and four may again contribute to this rather complex situation.

To Summarise, the effluent is implicated as the major vector leading to reduced limpet density near the discharge point. It is suggested that the effluent has the greatest affect on young age classes of limpet and establishment, through factors such as variation in effluent quality and rare imputs of emulsifier are possible contributary factors in removing some of the adult population.

The large size attained by the limpets near the outflow is mainly due to reduced infra and inter-specific competition accompanied by an abundant food supply though other factors contribute and are discussed.

The observed distributions at the central transects also differ from each other. The best explanation for this is offered by a recent hydrographic survey (Addy 1978) in which the effluent is seen to be retained around the effluent pipe itself (transect three) due to embayment of the discharge pipe. Thus, exposure to effluent material is even greater at transect three than at transect four. The survey also shows the effluent to be restricted to the
area around transects three and four at all stages of the tide.
If we also look at the results for the $P=0.05$ level of significance there are other differences that become apparent. The limpet distributions at transects two, five and six are significantly different from each other, though this is only border line significance, it is suggested that these observations are due to natural habitat variation to which P. vulgata is extremely sensitive (Blackmore 1969; Lewis and Bowman 1975). Transect six is one grade (Ballèntine 1961) more sheltered than transects two and five, and this might explain the greater median volume obtained. Transects five and six have suffered reductions in Balanus balanoides abundance since Crapp's 1970 survey, and Lewis and Bownan (I975) have snown a reduction in barnacle density to lead to increased growth rate and maximum attainable size in P. vulgatae This might also explain the greater limpet density found at transects five and six.
(ii) Quantitative Analysis of Petroleum Oil Pollutants in P. vulgata by
infra red spectrophotometry
The results aummarised in Table III show that the variation in both hydrocarbon Body Burden and Dry weight of limpet soft tissue, is greater between transects than within them. Transect three has a significantly higher hydrocarbon Body Burden than any other transect whereas transect four only differs significantly from transect one.

Dry weight determination also gave a higher mean value at transect three than at any other transect, with transect four differing from transects one, two, three and six but not transect five. Both the dry weight determination and the earlier size frequency investigation for $P_{\text {. vulgata }}^{2}$ show limpets at transects three to be significantly larger than at any other transect and for the trend to be repeated for transect four but to a lesser degreje. This is explained by differences in effluent retention between the two central transects: :The mean dry weight at transect five was not significantly different from that at transect four, and the reasons for this may be due to

Table III Results from the Statistical Analysis of the Epectrophotometry data given in Table II for P. vulgata Soft Tissue

| TECHNIQUE: Analysis of Variance (Model one ANDVAR) $Y_{i j}=\mu+Y_{i}+e i j$ | Analysis of Variance (Model one ANDVAR)$Y i j=\mu+\Psi_{i}+e i j$ |  |  |
| :---: | :---: | :---: | :---: |
| Data Source | Significance <br> Difference b | een means | Significantly different <br> transect means ( $P=0.05$ ) |
| $\begin{aligned} & C x \\ & (C x) \end{aligned}$ | $F_{5,48}=8.54$ | $\mathbf{P}=0.01$ | $T_{3}$ from all other transects $T_{4}$ from $T_{1}$ |
| Dry weight | $F_{5.48}=8.97$ | $\mathrm{P}=0.01$ | $T_{3}$ from all other transects $T_{4} \text { from } T_{1,2,3,6}$ |
| [] | $F_{5,48}=4.9$ | $P=0.01$ | $\begin{aligned} & T_{2} \text { from } T_{1,3,4,5} \\ & T_{6} \text { from } T_{1,3,4,5} \end{aligned}$ |

THCHNIQUE: Regression of $C x(\mathbb{Z})$ on Dry Wt. (X) of P. vulgata Soft Tissue

| Transect | Slope of Regression $t$ | Significance of Regression |  |
| :---: | :---: | :---: | :---: |
| All transects | 1.477 | $F_{1.52}=106.8$ | $\mathrm{P}=0.01$ |
| 1 | 1.658 | $F_{1,7}=28.91$ | $\mathrm{P}=0.01$ |
| 2 | 4.109 | $\mathrm{F}_{1,7}=21.54$ | $\mathbf{P}=0.01$ |
| 3 | 0.425 | $\mathrm{F}_{1,7}=0.88$ | N:S. |
| 4 | 1.529 | $F_{1,7}=9.28$ | $\mathbf{p}=0.05$ |
| 5 | 1.715 | $\mathrm{F}_{1,7}=34.9$ | $\mathbf{P}=0.01$ |
| 6 | 0.534 | $\mathrm{F}_{1,7}=4.21$ | N.S. |

Rable III (continued)

| TECHNIQOE: Comparison of the Slopes of the Significant Regressions ( $T_{1,2,4,5}$ ) with that of a slope of one |  |  |
| :---: | :---: | :---: |
| Transects | Slope | Significance of Difference between Regressions |
| All Significant <br> ones <br> 1 <br> 2 <br> 4 <br> 5 | $\begin{aligned} & 1.587 \\ & 1.658 \\ & 4.109 \\ & 1.529 \\ & 1.715 \end{aligned}$ | $\begin{aligned} & \mathbf{P}=0.01 \\ & \mathbf{P}=0.01 \\ & \mathbf{P}=0.01 \\ & \mathbf{P}=0.01 \\ & \mathbf{P}=0.01 \end{aligned}$ |

natural habitat variation effecting limpet size (Lewis and Bowman, 1975). One possible cause for the increased mean dry weight at transect five may be reduced interspecific competition, as Balanus balanoides has shown a decline in abundance since Crapp's survey in 1970. Although this may be a contributory factor it is insufficient to provide the sole explanation, since at transect six (subject to the same depletion in B. balanoides) a lower mean dry weight was found than at transect five over the population sampled.

When dry weight of a specimen was related to its body burden of hydrocarbon (hence providing an estimation of concentration (in an individual) no difference between the central transects and transects one and five were found. Pather surprisingly transects two and six held animals with a significantly higher mean concentration than the other transects. There appears to be no obvious explanation for such a result and the result for transect two must be viewed with some caution because the relation between Body Burden and Dry weight (see below) has a negative intercept which is rather suspect.

Since the limpets near the effluent discharge did not show higher mean hydrocarbon concentrations than those further away it was decided to investigate the relationship between Dry weight and Body Burden more fully. When the data were pooled irrespective of transect of origin, Body Burden was strongly dependant on Dry Weight (thus as Dry weight increased so did Body Burden) (figure thirteen). This relationship also held true for the individual transects (excluding transects three and six). The non significant result for these two transects may have arisen because the samples collected did not span a wide enough range of weights for a linear relationship to be detected.

Of rather more importance than the establishment of such a relationship, however, was the fact that the pooled data, and that for each individually significant regression, had slopes significantly greater than one. This indicates that as the weight of the limpets increased, so did the concentration of hydrocarbon in their soft tissue, irrespective of the transect of origin.


Figure Thirteen: Calculated Regression Lines of Body Burden vs. Dry Weight, for each transect and for all the data, pooled irrespective of Transect origin.

To summarise, two points of interest emerge from the spectrophotometry data:

1. That mean hydrocarbon concentration in P. vulgata is not greater near the effluent discharge.
2. That hydrocarbon concentration increases with increased Dry weight for transects one, two, four and five.

The first of these findings is open to a number of interpretations. Infra red spectrometry does not distinguish between hydrocarbona of Biogenic and Petrogenic origin and virtually all the hydrocarbon types found in crude oil occur naturally (Abus 1979, pers. comm.) Therefore, even with a "single point" analysis at the oil mixtures absorbtion maxima, Biogenic hydrocarbons will be present. The results might indicate firstly that all six transects de equaily coniaminatea or seconaly that there is differential contamination between the transects. Since the majority of the evidence here shows that contamination is far from evenly spread between the six transects, the first option must be extremely unlikely.

If the second alternative is accepted, there are a number of important deductions.

1. Infra red spectrometry has shown equal concentrations of hydrocarbons at the Little Wick Transects, these therefore are likely to be of Biogenic origin. It is not clear at present whether or not the marine invertebrates are synthesising their own hydrocarbons or whether their hydrocarbon content simply reflects the hydrocarbon content of their food source (Zsolnay et al, 1977). Lee et al (1977) have discovered that animals from areas of low, but constant petroleum input do not always show a markedly higher total hydrocarbon content relative to animals from cleaned areas and this leads on to the second deduction.
2. If Petrogenic hydrocarbons near the effluent outflow were ingested and accumulated by p. vulgata then this would show, over and above the Biogenic
concentrations found at all six transects. This is not the case and suggests that P. vulgata is either not taking the Petrogenic hydrocarbons into its system or that it has a mechanism for removing the Petrogenic hydrocarbons from its system once ingested. The first alternative is unlikely as an oil film was observed on the algae near the outflow and this is the limpets main food source. Also Teal (1976) has stated that it is clear from the chemistry of hydrocarbons that they should be absorbed through the guts of animals along with lipids in the diet. With the extra Petrogenic hydrocarbon present in the ambient water at certain stages of the tide, P. vulgata would almost inevitably absorb some oil into its system. Since ingestion of oil into the limpet system seems nigniy probaile tnen it is lukely that ro valgata has some depuration mechanism. This is not unlikely, various species of shrimp, crab's and lobster's rapidly take up petroleum by hydrocarbons from either the water or their food, (Anderson 1973; Cox et al. 1975; Sanborn and Malins 1976). Most of the hydrocarbons in the food were not assimilated by the tissues of the blue crab Callinectes sapidus but instead were immediately eliminated from the animal (Lee et al, 1976). Polychaetes: particulorly Gapibella capitata, are associated with areas of high oil input (Reish, 1971; Sanders et al, 1972). As a consequence, the polychaetes and quite possibly worms belonging to other phyla have evolved enzyme systems which metabolise Petroleum hydrocarbons (Lee 1976, Lee et al. 1977). Presumably hydrocarbon metabolism facilitates the rapid discharge of hydrocarbons observed for various species of Polychaetes.

The shore survey not only indicates that the limpet P. vulgata is in a contaminated environment at transects three and four, but it is likely that individuals live longer in this region of reduced recruitment and reduced intra and inter specific competition (Lewis and Bowman, 1975), thus they are subject to contamination for long periods of time. The fact that oil concentrations in their soft tissues are not significantly greater than of those limpets in
cleaner water makes the exist ${ }_{\text {ance }}$ of an efficient depuration mechanism highly likely.

The Second inference from this work, that hydrocarbon concentration increases with increased body weight, suggests that the larger limpets in any given area retain more hydrocarbon in their systems than do smaller ones, and Boyden (1977) has discovered a similar relationship for P. vulgata with Cadmium. Work on the landsnail Cepaea hortensis sampled at a suburban roadside showed that both total Cadmium and Cadmium concentrations increased with age for soft tissues, but at a given age Iarger animals had lower Cadmium concentrations than smaller ones (Williamson 1979). Lewis and Bowman (1975) have shown that the accurate age determination of limpets is extremely uncertain (especially if comparisons are to be made between different sites) so to investigate the differential effects of size and age without observing a cohort of P. vulgata from settlement to maturity would be unreliable, but as an idea for a future investigation it is very promising.

Another important topic for future study would be to differentiate between Biogenic and Detrogenic hydrocarbons in limpet tiscue from the six Little Wick transects. This could best be achieved by Gas Liquid Chromatography (Zsolnay, 2977), comparing traces for extraction solutions from limpet samples (as obtained by the above method) with a sample of the effluent itself. The analysis of some fresh limpet material from a known unpolluted site would be a valuable addition to the study as well.
A (1) Shore Survey Data for 1979 Survey ..... 53Archer-Thomson, J.H.S.
A (2) Shore Survey Data for 1978 Survey ..... 60
Petpiroon, S.
A (3) Shore Survey Data for 1970 Survey ..... 67
Crapp, G.B.
A (4) Criteria of Abundance for Common Plants and ..... 74
Animals of Rocky Sea Shores
 ..... 77
for Little Wick Transect Five
A (6) Raw Data for Limpet Volume Index ..... 79
A (7) Details of the mean effluent quality discharged ..... 93
in Little Wick Bay from the Esso Refinery

A (1) Shore Survey Data. Archer-Thomson, J.H.S. (1979)

Key (for all Shore Survey Data)

1 - Rare (R)
2 - Occasional
(0)

3 - Frequent (F)
4 - Common (C)
5 - Abundant (A)
6 - Super Abundant (s)
7 - Extremely Abundent (Ex)

D - In dead barnacles
C - In crevices
P - Present but outside survey area
S - On seaweed
U - Understones

* = In Rock Pool



サT $\mathcal{T}$ 工T TL OT $68<9 \leq \pm \leq$




A (2) Shore Survey Data. Petpiroon, S. (1978)




|  |  |  |  |  |  |  |  |  |  |  | $\operatorname{st\text {tape}} \frac{\operatorname{sntit} \Omega_{W}}{\operatorname{s\tau \partial ssn_{W}}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  | sntiṭdet eqtoonn |
|  |  | $\varepsilon$ $\downarrow$ |  |  |  |  | $\left[\begin{array}{c} s^{2} \\ a^{7} \\ 0^{\circ} \\ 0^{2} E \\ \end{array}\right.$ | E $E$ |  |  |  |
|  |  |  |  |  |  |  | 1 |  | ＋ | ! |  <br> Eェエexeuto einqsio streotrtqun eqnqqio SITə पsdat |
|  |  |  |  | 27 | 7 ¢ | $E d^{\text {c }}$ | $\varepsilon$ | $7{ }^{7} 8$ |  |  | $\text { ds } \frac{\text { erfazed }}{\text { sfadutit }}$ |
|  |  |  | 2 |  | 2  <br> 7 7 <br> 2 2 |  | $\varepsilon$  <br> $\varepsilon$ $\varepsilon$ <br>   <br>   |  |  | 2 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | $\qquad$ | $1$ |  |  |  | ， | 12 | $21^{6}$ |  |  |


WORKERR S.P.



A (3) Shore Survey Data. Crapp, G.B. (1970)








A (4) Criteria of Abundance for Common Plants + Animals of Rocky Sea Shores

Live barnacles (except B. perforatus)
(record adults, spat, cyprides sep'tly)

1. Littorina neritoides

## Littorina neglecta

Ex 500 or more per $0.01 \mathrm{~m}^{2}-5+/ \mathrm{cm}^{2}$
S 300-499 per $0.01 \mathrm{~m}^{2} \quad 3-4 / \mathrm{cm}^{2}$
A $100-299$ per $0.01 \mathrm{~m}^{2} \quad 1-2 / \mathrm{cm}^{2}$
C 10-99 per $0.01 \mathrm{~m}^{2}$
F 1-9 per 0.01
0 1-99 per $\mathrm{m}^{2}$
2. Balanus perforatus

Ex 300 or more per $0.01 m^{2}$
s 100-299 per 0.01m ${ }^{2}$
A $10-99$ per $0.01 m^{2}$
C 1-9 per $0.01 \mathrm{~m}^{2}$
F $1-9$ per $0.1 m^{2}$
0 1-9 per $m^{2}$
R Less than 1 per $m^{2}$
3. Patella spp. $10 \mathrm{~mm}+$

Littorina littorea (juvs \& ads)
Littorina littoralis (adults)
juv. Nucella lapillus ( 3 mm )
Bx 20 or more per $0.1 m^{2}$
S $10-19$ per $0.1 m^{2}$
A 5-9 per $0.1 m^{2}$
C $1-4$ per $0.1 m^{2}$
3. continued

F $\quad 5-9$ per $\mathrm{m}^{2}$
$0 \quad 1-4$ per $m^{2}$
R Less than 1 per $m^{2}$
4. Iittorina 'saxatilis'

Patella smaller than 10 mm
Anurida maritima

Hyale nilssoni \& other amphipods
juvenile I. Iittoralis
Ex 50 or more per $0.1 m^{\text {a }}$
S 20-49 per $0.1 \mathrm{~m}^{2}$
A 10-19 per $0.1 \mathrm{~m}^{2}$
C $5-9$ per $0.1 \mathrm{~m}^{2}$
F $\quad 1-4$ per $0.1 \mathrm{~m}^{2}$
0 1-9 per $m^{2}$
$F$ Less than 1 per min $^{2}$
5. Nucella lapillus ( 3 mm )

Gibbula spp., Monodonta lineata

## Actinea equina

Idotea granulosa
Juv. \& recent sett. Carcinus

## Iigea oceanica

Ex 10 or more per $0.1 m^{2}$
S 5-9 per $0.1 \mathrm{~m}^{2}$
A $1-4$ per $0.1 \mathrm{~m}^{2}$
5. continued

C 5-9 per $\mathrm{m}^{2}$ locally sometimes more
F 1-4 per m ${ }^{2}$ locally sometimes more
0 less than 1 per $\mathrm{m}^{2}$ locally sometimes more
R Always less than 1 per $\mathrm{m}^{2}$
6. Mytilus edulis

Ex $80 \%$ or more cover
S 50-79\% cover
A 20-49\% cover
C $\quad 5-19 \%$ cover
F Small patches $5 \% \ldots 10+$ sm. inds. per $0.1 \mathrm{~m}^{2}$, 1 or more lg . per $0 . \mathrm{Im}^{2}$
0 1-9 sm. per $0.1 \mathrm{~m}^{2}, 1-9 \mathrm{lg}$. per $\mathrm{m}^{2}$ No patches except sm. in crevices
$R$ Less than 1 per $\mathrm{m}^{2}$
7. Fomatoceros triqueter

A 50 or more tubes per $0.01 \mathrm{~m}^{2}$
C 1-49 tubes per $0.01 \mathrm{~m}^{2}$
F 1-9 tubes per $0.1 \mathrm{~m}^{2}$
0 1-9 tubes per $\mathrm{m}^{2}$
$R$ Less than 1 tibe per $m^{2}$
8. Spirobis spp.

A 5 or more per $\mathrm{cm}^{2}$ on approp. substs. More than 100 per $0.01 \mathrm{~m}^{2}$ generally
C Patches of 5 or more per $\mathrm{cm}^{2}$ $1-100$ per $0.01 \mathrm{~m}^{2}$ generally

F Widely scattered small groups $1-9$ per $0.1 \mathrm{~m}^{2}$ generally

## 8.continued

0 Widely scattered small groups
Less than 1 per $0.1 \mathrm{~m}^{2}$ generally
$R$ Less than 1 per $\mathrm{m}^{2}$

## 9. Sponges

Hydroids
Bryozoa
A Present on $20 \%$ or more suit. surf.

C Present on 5-19\% of suit, surf.

F Scattered patches, less than 5\% cover

0 Small patch or single Sprig in $0.1 \mathrm{~m}^{2}$
$R$ Less than one patch over strip, one amall patch or sprig per $0.1 \mathrm{~m}^{2}$
10. Flowering Plants, lichens \& 1ithothamia

Ex More than $80 \%$ cover
S 50-79\% cover
A $\mathbf{2 0 - 4 9 \%}$ cover
C 1-19\% cover
F Large scattered patches
0 Widely scattered patches, all small

R Only 1 or 2 patches
11. Algae

Ex More than $90 \%$ cover
S 60-89\% cover
A 30-59\% cover
C 5-29\% cover
F Less than 5\% cover, zone still apparent
0 Scattered plants, zone indistinct
R Only l or 2 plants
Other animal species
Record as \% cover or approx. average
numbers within $0.01,0.1$ or $\mathrm{lm}^{2}$

A (5) Biological Exposure Score Sheet (Ballantine, 1961) for Little Wick Transect 5

| $N=$ Absent | $R=$ Rare | $O=$ Occasional |
| :--- | :--- | :--- |
| $F=$ Frequent | $C=$ Common | $A=$ Abundant |


|  | Exposed |  |  | : |  | Sheltered |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Exposure Grade | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| Alaria esculenta | A | $\mathrm{N}-\mathrm{A}$ | N-A | $\mathrm{N}-\mathrm{A}$ | N | N | N |
| Porphyra umbilicalis (hlf) | A | C-A | N-A | N | N | N | N |
| Laminaria hyperborea | N-D | O-A | R-C | N | N | N | N |
| Lichina pygmaea | $\mathrm{N}-\mathrm{C}$ | C-A | C-A | R-A | N | N | N |
| Mytilus edulis | R-A | R-A | D-1 | D_T | D-T | M!-n | *! |
| Corallina officinalis | A | F-A | F-A | O-C | O-C | $\mathrm{N}-\mathrm{F}$ | N |
| Patella depressa | O-C | O-A | O-A | O-A | O-A | $\mathrm{N}-\mathrm{C}$ | N |
| Patella aspera | F-A | F-A | F-A | R-A | R-A | N-A | N |
| Littorina neritoidea | C-A | C-A | C-A | C-A | C-A | R-C | N |
| Laurencia spp. | O-F | O-A | O-A | O-A | O-A | R-A | N |
| Chthamalus stellatus | A | A | A | C-A | C-A | C-A | R-F |
| Fucus serratus | N | R-A | $\mathrm{R}-\mathrm{A}$ | R-A | R-A | A | A |
| Balanus perforatus | N | N-F | O-A | O-A | O-A | O-A | O-A |
| Pelvetia canaliculata | N | N | R-F | R-F | F-C | A | A |
| Nucella lapillus | N | iv | $\boldsymbol{R}-\mathbf{A}$ | C-A | C-A | C-A | C |
| Spirobis rupestrus | N | N | $\mathrm{N}-\mathrm{O}$ | O-F | $\mathrm{C}-\mathrm{A}$ | C-A | C-A |
| Gibbula umbilicalis | N | $\mathrm{N}-\mathrm{A}$ | $\mathrm{N}-\mathrm{A}$ | C-A | A | A | A |
| Littorina obtusata | N | N | N | R-F | R-F | F-A | A |
| Littorina littorea | $\mathbf{N}$ | N | N | N-C | R-C | R-C | A |



$$
\text { Exposure Grade } 4
$$

Date surveyed: 12.5.79

A (6) Raw Data for P. vulgata Volume - Index

Eey

$$
\begin{aligned}
\longleftrightarrow & =\text { Longest Diameter (mm) } \\
\uparrow & =\text { Shortest Diameter (mm) } \\
\Delta & =\text { Height of Shell (mm) } \\
V & =\text { Volume }\left(\mathrm{cm}^{3}\right)
\end{aligned}
$$

Transect One: 28.6.79

| $\longleftrightarrow$ | $\uparrow$ | $\wedge$ | $V$ |
| :---: | :---: | :---: | :---: |
| 27 | 22 | 13 | 2.042 |
| 23 | 20 | 12 | 1.458 |
| 24 | 18 | 6 | 0.691 |
| 10 | 6 | 4 | 0.067 |
| 20 | 14 | 6 | 0.452 |
| 20 | 18 | 12 | 1.131 |
| 21 | 16 | 9 | 0.811 |
| 20 | 15 | 9 | 0.726 |
| 5 | 4 | 2 | 0.010 |
| 12 | 10 | 6 | 0.188 |
| 13 | 11 | 6 | 0.226 |
| 21 | 14 | 11 | 0.889 |
| 12 | 8 | 4 | 0.105 |
| 13 | 8 | 4 | 0.117 |
| 13 | 10 | 5 | 0.173 |
| 12 | 10 | 4 | 0.126 |
| 12 | 9 | 4 | 0.117 |
| 12 | 8 | 6 | 0.157 |
| 19 | 12 | 9 | 0.565 |
| 12 | 8 | 7 | 0.183 |
| 13 | 10 | 6 | 0.207 |
| 12 | 10 | 5 | 0.157 |
| 16 | 11 | 4 | 0.193 |
| 20 | 15 | 10 | 0.806 |
| 5 | 4 | 3 | 0.016 |
| 15 | 10 | 5 | 0.204 |
| 20 | 18 | 12 | 1.131 |
| 14 | 10 | 6 | 0.226 |
| 20 | 16 | 9 | 0.263 |
| 26 | 22 | 14 | 2.111 |
| 25 | 20 | 11 | 1.463 |
| 14 | 11 | 7 | 0.286 |
| 10 | 8 | 4 | 0.084 |
| 10 | 8 | 5 | 0.105 |

Sample Point : Station 4 Low water $=15.32$

Height 1.3 m

Area Sampled $=\frac{1}{2}$ a square metre

Transect One (continued)

| $\longleftrightarrow$ | $\downarrow$ | $\Delta$ | $V$ |
| :---: | :---: | :---: | :---: |
| 14 | 12 | 6 | 0.264 |
| 12 | 10 | 5 | 0.157 |
| 13 | 11 | 5 | 0.188 |
| 10 | 9 | 4 | 0.096 |
| 12 | 10 | 4 | 0.126 |
| 12 | 8 | 4 | 0.105 |
| 17 | 12 | 9 | 0.499 |
| 26 | 20 | 12 | 1.659 |
| 10 | 7 | 4 | 0.075 |
| 24 | 20 | 10 | 1.267 |
| 23 | 20 | 10 | 1.215 |
| 14 | 10 | 5 | 0.916 |
| 18 | 12 | 8 | 0.469 |
| 18 | 14 | 6 | 0.402 |
| 14 | 12 | 7 | 0.308 |
| 10 | 8 | 4 | 0.084 |
| 12 | 9 | 4 | 0.117 |
| 21 | 19 | 3 | 1. 367 |
| 12 | 9 | 7 | 0.205 |
| 14 | 10 | 6 | 0.226 |
| 13 | 10 | 5 | 0.173 |
| 16 | 12 | 5 | 0.257 |
| 15 | 10 | 4 | 0.163 |
| 16 | 12 | 6 | 0.308 |
| 18 | 16 | 7 | 0.528 |
| 16 | 12 | 7 | 0.359 |
| 20 | 16 | 10 | 0.848 |
| 30 | 22 | 17 | 3.009 |
| 21 | 16 | 10 | 0.901 |
| 18 | 16 | 10 | 0.890 |
| 30 | 20 | 12 | 1.960 |
| 20 | 15 | 9 | 0.726 |
| 30 | 26 | 15 | 3.079 |
| 18 | 14 | 7 | 0.469 |
| 26 | 12 | 20 | 1.885 |
| 26 | 22 | 14 | 2.111 |
| 12 | 10 | 4 | 0.126 |
| 24 | 16 | 15 | 1.571 |
| 25 | 16 | 14 | 1.579 |
| 14 | 12 |  | 0.396 |
| 16 | 15 |  | 0.565 |

$$
\mathrm{n}=75
$$

| $\longleftrightarrow$ | $\downarrow$ | $\Delta$ | $V$ |
| ---: | ---: | ---: | ---: |
| 20 | 13 | 8 | 0.570 |
| 9 | 6 | 4 | 0.059 |
| 20 | 16 | 9 | 0.763 |
| 16 | 10 | 7 | 0.308 |
| 15 | 10 | 7 | 0.286 |
| 22 | 16 | 8 | 0.754 |
| 16 | 10 | 9 | 0.396 |
| 20 | 14 | 7 | 0.528 |
| 25 | 18 | 14 | 1.701 |
| 14 | 12 | 6 | 0.264 |
| 18 | 15 | 10 | 0.712 |
| 14 | 10 | 6 | 0.226 |
| 19 | 14 | 10 | 0.712 |
| 20 | 16 | 8 | 0.679 |
| 23 | 19 | 14 | 1.466 |
| 20 | 16 | 9 | 0.763 |
| 16 | 14 | 6 | 1.097 |
| 20 | 14 | 9 | 0.679 |
| 10 | 8 | 5 | 0.105 |
| 19 | 13 | 8 | 0.536 |
| 10 | 7 | 4 | 0.075 |
| 16 | 14 | 6 | 0.352 |
| 14 | 10 | 5 | 0.183 |
| 13 | 10 | 5 | 0.356 |
| 14 | 10 | 4 | 0.151 |
| 15 | 11 | 6 | 0.264 |
| 16 | 12 | 5 | 0.257 |
| 17 | 14 | 5 | 0.314 |
| 20 | 15 | 9 | 0.726 |
| 12 | 10 | 4 | 0.126 |
| 22 | 14 | 11 | 0.933 |
| 20 | 15 | 10 | 0.806 |
| 16 | 13 | 6 | 0.320 |
| 18 | 14 | 10 | 0.670 |
| 18 | 11 | 8 | 0.444 |
| 18 | 13 | 10 | 0.628 |
| 14 | 11 | 6 | 0.245 |
| 11 | 6 | 3 | 0.057 |
| 10 | 6 | 3 | 0.050 |
| 17 | 12 | 10 | 0.555 |
| 26 | 22 | 14 | 2.111 |
| 21 | 18 | 11 | 1.094 |
| 15 | 12 | 5 | 0.241 |
| 6 | 4 | 3 | 0.019 |
| 26 | 20 | 14 | 1.935 |
| 14 | 11 | 5 | 0.204 |
| 20 | 14 | 6 | 0.452 |
| 5 | 4 | 3 | 0.016 |
|  |  |  |  |

Sample Point : Station 4
Low water $=4.04$
$H_{\text {eight }}=1.5 \mathrm{~m}$

Area sampled $=\frac{1}{2}$ square metre

Transect Two (continued)

| $\longleftrightarrow$ | $\uparrow$ |  |  |
| ---: | ---: | ---: | :---: |
| 10 | 8 | 4 | 0.084 |
| 8 | 6 | 4 | 0.050 |
| 16 | 10 | 6 | 0.264 |
| 24 | 20 | 14 | 1.774 |
| 16 | 10 | 6 | 0.264 |
| 14 | 10 | 5 | 0.188 |
| 18 | 14 | 6 | 0.402 |
| 17 | 14 | 6 | 0.377 |
| 20 | 15 | 6 | 0.484 |
| 17 | 13 | 10 | 0.586 |
| 18 | 13 | 10 | 0.670 |
| 16 | 12 | 9 | 0.462 |
| 16 | 14 | 6 | 0.352 |
| 21 | 18 | 14 | 1.393 |
| 18 | 14 | 6 | 0.402 |
| 18 | 13 | 10 | 0.628 |
| 18 | 16 | 6 | 0.452 |
| 9 | 6 | 2 | 0.044 |
| 15 | 12 | 8 | 0.385 |
| 19 | 17 | 12 | 1.018 |
| 10 | 6 | 5 | 0.084 |
| 14 | 12 | 8 | 0.352 |
| 15 | 12 | 9 | 0.434 |
| 10 | 7 | 6 | 0.113 |
| 19 | 14 | 8 | 0.569 |
| 22 | 18 | 14 | 1.466 |
| 15 | 10 | 5 | 0.204 |
| 16 | 10 | 6 | 0.264 |
| 16 | 12 | 6 | 0.308 |
| 10 | 7 | 3 | 0.057 |
| 14 | 10 | 6 | 0.226 |
| 20 | 14 | 7 | 0.528 |
| 10 | 6 | 4 | 0.067 |
| 9 | 6 | 3 | 0.044 |
| 20 | 15 | 9 | 0.726 |
| 17 | 10 | 6 | 0.289 |
| 16 | 12 | 9 | 0.462 |
| 14 | 12 | 6 | 0.264 |
| 15 | 12 | 7 | 0.337 |
| 14 | 10 | 6 | 0.226 |
| 24 | 18 | 10 | 1.152 |

$$
n=89
$$

|  | $\uparrow$ |
| :---: | :---: |
|  | $\leftrightarrow$ |
|  | $D$ |
|  | $<$ |

Sample Point : Station 4
Low Water $=4.04$
Height $=1.5 \mathrm{~m}$

Area Sampled $=\frac{1}{2}$ square meter

## Transect Three (continued)

| $\longleftrightarrow$ | $\uparrow$ | $\Delta$ | $\vee$ |
| ---: | ---: | ---: | ---: |
| 25 | 17 | 9 | 1.037 |
| 18 | 16 | 7 | 0.528 |
| 16 | 13 | 5 | 0.278 |
| 60 | 51 | 35 | 28.222 |
| 38 | 30 | 19 | 5.750 |
| 15 | 12 | 4 | 0.193 |
| 38 | 32 | 19 | 6.088 |
| 40 | 36 | 20 | 7.561 |
| 41 | 37 | 20 | 7.77 |
| 36 | 31 | 16 | 4.708 |
| 38 | 32 | 17 | 5.448 |
| 42 | 38 | 21 | 8.796 |
| 46 | 37 | 25 | 11.284 |
| 52 | 43 | 31 | 18.309 |
| 34 | 28 | 16 | 4.021 |

$$
n=61
$$

| $\leftrightarrow$ | $\uparrow$ | $\Lambda$ | $V$ |
| :---: | ---: | ---: | ---: |
| 37 | 30 | 21 | 6.179 |
| 40 | 36 | 18 | 6.805 |
| 34 | 28 | 18 | 4.524 |
| 38 | 36 | 19 | 6.805 |
| 51 | 40 | 25 | 13.561 |
| 36 | 26 | 15 | 3.769 |
| 28 | 21 | 12 | 1.885 |
| 21 | 18 | 9 | 0.896 |
| 34 | 29 | 17 | 4.415 |
| 33 | 28 | 15 | 3.659 |
| 36 | 30 | 16 | 4.557 |
| 49 | 37 | 27 | 13.063 |
| 30 | 25 | 13 | 2.573 |
| 44 | 35 | 17 | 6.943 |
| 47 | 34 | 21 | 9.016 |
| 48 | 38 | 20 | 9.676 |
| 53 | 47 | 27 | 17.671 |
| 18 | 14 | 7 | 0.469 |
| 37 | 26 | 20 | 5.194 |
| 25 | 21 | 14 | 1.935 |
| 32 | 23 | 20 | 3.958 |
| 16 | 14 | 5 | 0.293 |
| 60 | 49 | 35 | 27.232 |
| 30 | 22 | 15 | 2.655 |
| 17 | 13 | 7 | 0.411 |
| 16 | 13 | 5 | 0.278 |
| 13 | 10 | 4 | 0.138 |
| 16 | 12 | 5 | 0.257 |
| 23 | 17 | 10 | 1.047 |
| 18 | 14 | 9 | 0.603 |
| 19 | 14 | 10 | 0.712 |
| 22 | 20 | 9 | 1.037 |
| 20 | 16 | 11 | 0.933 |
| 39 | 32 | 18 | 5.938 |
| 28 | 34 | 13 | 3.267 |
| 17 | 14 | 9 | 0.565 |
| 25 | 23 | 13 | 1.960 |
| 40 | 30 | 21 | 6.729 |
| 15 | 11 | 7 | 0.308 |
| 20 | 16 | 17 | 1.442 |
| 36 | 30 | 16 | 4.557 |
| 27 | 22 | 12 | 1.885 |
| 23 | 22 | 11 | 1.463 |
| 36 | 30 | 18 | 5.127 |
| 35 | 28 | 20 | 5.194 |
| 16 | 12 | 5 | 0.257 |
| 32 | 28 | 15 | 3.534 |
| 25 | 20 | 11 | 1.463 |
|  |  |  |  |

Sample Point : Station 4<br>Low water $=4.04$<br>Height $=1.5 \mathrm{~m}$<br>Area sampled $=\frac{1}{2}$ square metre

Transect Four (continued)

| $\leftrightarrow$ | $\uparrow$ | $\Delta$ | $V$ |
| ---: | ---: | ---: | ---: |
| 35 | 30 | 15 | 4.147 |
| 20 | 14 | 9 | 0.679 |
| 34 | 30 | 15 | 4.021 |
| 20 | 13 | 28 | 0.641 |
| 30 | 23 | 7 | 2.765 |
| 50 | 42 | 10 | 15.511 |
| 16 | 13 | 6 | 0.389 |
| 25 | 23 | 9 | 1.508 |
| 18 | 13 | 8 | 0.377 |
| 27 | 20 | 9 | 1.301 |
| 16 | 13 | 8 | 0.444 |
| 39 | 31 | 21 | 6.729 |
| 30 | 29 | 15 | 3.424 |
| 47 | 40 | 29 | 14.364 |

$$
\mathrm{n}=62
$$

|  | $\downarrow$ |
| :---: | :---: |
|  | $\leftrightarrow$ |
|  | $D$ |
|  | $<$ |

```
Sample Point: Station 4
Low water \(=18.11\)
Height \(=2.2 \mathrm{~m}\)
Area sampled \(=\frac{1}{2}\) square metre
```

```
Transect Five (continued)
```

|  | $\uparrow$ |  |  |
| ---: | ---: | ---: | :---: |
| 9 | $\downarrow$ | $\Delta$ | $V$ |
| 11 | 8 | 4 | 0.075 |
| 15 | 10 | 5 | 0.120 |
| 20 | 15 | 9 | 0.286 |
| 11 | 8 | 4 | 0.726 |
| 20 | 16 | 10 | 0.096 |
| 20 | 16 | 10 | 0.848 |
| 17 | 11 | 9 | 0.462 |
| 21 | 16 | 9 | 0.811 |
| 14 | 10 | 6 | 0.226 |
| 21 | 15 | 10 | 0.440 |
| 17 | 13 | 7 | 0.411 |
| 16 | 12 | 7 | 0.359 |
| 16 | 14 | 10 | 0.586 |
| 17 | 12 | 10 | 0.555 |
| 6 | 4 | 3 | 0.019 |
| 8 | 7 | 1 | 0.052 |
| 13 | 8 | 6 | 0.176 |
| 18 | 13 | 5 | 0.314 |
| 13 | 9 | 4 | 0.126 |
| 12 | 8 | 5 | 0.131 |
| 20 | 16 | 10 | 0.848 |
| 18 | 12 | 5 | 0.293 |
| 18 | 13 | 7 | 0.440 |
| 17 | 14 | 9 | 0.565 |
| 11 | 8 | 7 | 0.169 |
| 9 | 6 | 4 | 0.059 |
| 15 | 12 | 7 | 0.357 |
| 13 | 11 | 7 | 0.264 |
| 20 | 14 | 10 | 0.754 |
| 24 | 20 | 9 | 1.140 |
| 25 | 22 | 15 | 2.168 |
| 19 | 18 | 10 | 0.901 |
| 22 | 16 | 9 | 0.848 |
| 18 | 10 | 9 | 0.462 |
| 10 | 7 | 4 | 0.075 |
| 11 | 8 | 5 | 0.120 |
| 18 | 12 | 8 | 0.469 |
| 14 | 11 | 7 | 0.264 |
| 14 | 10 | 5 | 0.188 |
| 16 | 15 | 7 | 0.440 |
| 16 | 14 | 10 | 0.586 |
| 16 | 12 | 7 | 0.359 |
| 17 | 15 | 8 | 0.536 |
| 23 | 19 | 11 | 1.267 |
| 15 | 11 | 6 | 0.264 |
| 14 | 10 | 4 | 0.151 |
| 16 | 12 | 8 | 0.411 |
| 19 | 15 | 10 | 0.754 |
| 12 | 9 | 4 | 0.117 |
| 15 | 8 | 5 | 0.173 |
|  |  |  |  |


| $\leftrightarrows$ | $\downarrow$ | $\Delta$ | $\vee$ |
| :---: | ---: | ---: | :---: |
| 16 | 12 | 5 | 0.257 |
| 20 | 14 | 11 | 0.829 |
| 22 | 18 | 10 | 1.047 |
| 24 | 20 | 13 | 1.647 |
| 18 | 14 | 9 | 0.603 |
| 23 | 20 | 10 | 1.215 |
| 23 | 20 | 13 | 1.579 |
| 18 | 12 | 6 | 0.352 |
| 20 | 16 | 11 | 0.933 |
| 15 | 10 | 7 | 0.264 |
| 24 | 20 | 12 | 0.452 |
| 11 | 9 | 4 | 0.105 |
| 20 | 15 | 10 | 0.806 |
| 24 | 18 | 10 | 1.152 |
| 18 | 14 | 11 | 0.737 |
| 15 | 10 | 6 | 0.245 |
| 12 | 7 | 5 | $0.11 . r$ |
| 17 | 13 | 8 | 0.469 |
| 11 | 7 | 4 | 0.084 |
| 12 | 10 | 7 | 0.220 |
| 15 | 11 | 6 | 0.264 |
| 18 | 16 | 11 | 0.829 |

$$
\mathrm{n}=118
$$

|  | $\downarrow$ |
| :---: | :---: |
|  | $\longleftrightarrow$ |
|  | $D$ |
|  <br>  | $<$ |

Sample Point : Station 4
Iow Tide $=18.11$
Height $=2.2 \mathrm{~m}$

Area sampled $=\frac{1}{2}$ square metre

Transect Six (continued)

|  | $\downarrow$ |
| :---: | :---: |
|  | $\leftrightarrow$ |
|  | $D$ |
| OHトOOOOOOOOHOOOOOOOHトHOHOOOHOCOHOOHOOOOHOOOOOOOO <br>  | $<$ |

Transect Six (continued)

| $\longleftrightarrow$ | $\downarrow$ | $\Delta$ | $V$ |
| :---: | ---: | ---: | :---: |
| 11 | 9 | 6 | 0.157 |
| 20 | 15 | 13 | 1.048 |
| 13 | 10 | 9 | 0.311 |
| 16 | 12 | 9 | 0.462 |
| 20 | 15 | 12 | 0.967 |
| 20 | 17 | 15 | 1.351 |
| 12 | 9 | 7 | 0.205 |
| 17 | 14 | 9 | 0.763 |
| 10 | 8 | 6 | 0.126 |
| 17 | 15 | 8 | 0.536 |
| 19 | 16 | 9 | 0.726 |
| 26 | 20 | 16 | 2.212 |
| 19 | 16 | 10 | 0.806 |
| 18 | 14 | 8 | 0.536 |

$\mathrm{n}=111$

A (7) Details of the mean Effluent Quality discharged in Little Wick Bay from the Esso Refinery

Mean Effluent Quality in 1970

| Oil content | - | $25 \mathrm{mg} / \mathrm{l}$ |
| :--- | :---: | :---: |
| pH | - | 8.0 |
| Temperature | - | $80^{\circ} \mathrm{F}$ |
| Phenols | - | $0.3 \mathrm{mg} / \mathrm{l}$ |
| Suspended Solids | - | $50 \mathrm{mg} / \mathrm{l}$ |
| Oxygen absorbed from | - | $10.2 \mathrm{mg} / \mathrm{I}$ |
| $\quad$ acid permangenate |  |  |
| $\mathrm{H}_{2} \mathrm{~S}$ | - | Not detectable |
| $\mathrm{NH}_{3}$ | - | $1.5 \mathrm{mg} / \mathrm{l}$ |

Effluent quality has not changed significantly since 1970 except for the fact that oil content has reduced from a mean of $25 \mathrm{mg} / 1$ to one of $15 \mathrm{mg} / 1$.

Information supplied by the Fiss Petroleum Company, Iimited, Milford Haven, 1979。

## REFHRTNCES

1. ADDY, J.M. (1978) Biollogical and hydrographic survey off Little Wick Bay, Milford Haven, O.P.R.U. Annual Report 1977/1978.
2. ADMIRALTY HYDROGRAPHIC DEPARTMENT. The Admiralty Tide Tables (1) Buropean Waters, H.M.S.O. London.
3. ANDERSON, J.W. (1975) Laboratory Studies on the effects of oil on marine organisms: an overview, American Petroleum Institute Publication, No. 4349, American Petroleum Institute, Washington, D.C., 1975
4. BAKER, J.M. (1975) Investigation of refinery effluent effects through field surveys, in Marine Ecology and Oil Pollution (ed. J.M. Baker), Appiied Science Fubilishers, for Institute of Petroleum, Essex, England, 201.-225
5. BAKBR, J.M. (1975a) Biological monitoring - principles, methods and difficulties, in Marine Ecology and Oil Pollution (ed. J.M. Baker).
6. BAKER, J.M. (1975b) Experimental investigation of refinery effluents, in Marine Ecology and Oil Pollution (ed. J.M. Baker).
7. BATLANTINE, W.J. (1961). A Biologically-defined exposure scale for the comparative description of rocky shores. Field Studies, I (3), 1-19
8. BARREIT, J. and YOUNGE, C.M. (1958). Collins Pocket Guide to the sea shore, Collins, Iondon.
9. BLACKMORE, D.T. (1969) Studies of Patella vulgata L.1. Growth, reproduction and zonal distribution. J. exp. mar. Biol. Ecol., Vol. 3, pp 200-213
10. BOYDEN, P. (1977) Effect of size upon metal content of shellfish. J. mar. biol. Ass. U.K. 57, 675-714
11. BRATMTEGARD, T. (1966) The natural history of the Wardangerfjord. 7. Horizontal distribution of the fauna of rocky shores, Sarsia, 22, 1-54
12. COX, B.A., ANDERSON, J.W. and PARKER, J.C. (1975). An experimental oil spill: The distribution of aromatic hydrocarbons in the water, sediment, and animal tissues within a shrimp pond. pp 607-612. In: Proceedings of the Conference on Prevention and Control of Oil Pollution, American Petroleum Institute, Washington, D.C., 1975.
13. CRAPP, G.B. (1970). The Biological effects of Marine Oil Pollution and Shore Cleansing, Ph.D. Thesis. University of Wales, Swansea.
14. CRISP, D.J. and SOUTHWARD, A.J. (1958). The distribution of intertidal organisms along the coasts of the English Channel. J. mar. biol. Ass. U.K. 37, 157-208
15. DICKS, B. (1973). Some effects of Kuwait crude oil on the limpet, Patella vulgata. miviron. Mollut. 2, 219-22y
16. DICKS, B. (1975b) The importance of behavioural patterns in toxicity testing and ecological prediction in Marine Ecology and Oil Pollution (ed. J.M. Baker).
17. DUDLEX, M. (1968). Oil Pollution in Milford Haven, Field Studies, 2 (Suppl), 21-29.
18. EVANS; R.G. (1947), The intertidal ecology of rocky shores in the Plymouth neighbourhood. J. mar. biol. Ass. U.K., 27, 173-218
19. GEORGE, M. (1961). Oil pollution of marine organisms. Nature, Iond. 192, 1209
20. GRUENFKD, M. (1975) "Quantitative analysis of Petroleum Oil Pollutants by Infrared spectrophotometry" Water Quality Parameters, ASTM STP 573, American society for testing and materials, pp 290-308
21. HARVA, O. and SOMERSALD, A. (1958). Suomen Kemistilehti, 31 (b) pp 384-387: Cited in Gruenfeld (1975).
22. KNIGHT-JONES, E.W. (1953b). Laboratory experiments on gregariousness during setting in Balanus balanoides and other barnacles. J. exp. biol. 30, 584-598
23. LEE, R.F. (1976). Metabolism of petroleum hydrocarbons in marine sediments, pp 334-344. In: Sources, Effects and Sinks of Petroleum in the Aquatic Fnvironment, American Institute of Biological Sciences, Washington, D.C.
24. LEEE, R.F., RYAN, C. and NEUHAUSER, M.I. (1976). Fate of petroleum hydrocarbons taken up from food and water by the blue crab, Callinectes sapidus, Mar. Biol. 37, 363-370
25. LEE, R.F., FURIONG, F. and SINGER, S. (1977). Metabolism of hydrocarbons hydroxylase from the tissues of the blue crab, Callinectes sapidus, and the Polychaete worm, Neries spp. In: C.S. Giam (ed) Pollutant Effects of Marine Organisms, D. $\overline{\text { G }}$. Heath, Lexington, Massachusetis.
26. LEAIS, J.R. (1964). The Ecology of Rocky Shores
27. LIWIS, J.R. and BOWMAN, R.S. (1975). Local Habit-induced variations in the population dynamics of Patella vulgata I. J. exp. mar. Biol. Ecol., 17 pp 165-203.
28. MIIFORD HAVEN CONSERVANCY BOARD (1968). The Port of Milford Haven, Milford Haven Conservancy Board, pp. 32.
29. MOYSE, J. and NESSON-SMIIM, S. (1963). Zonation of animals and plants on rocky shores around Dale, Pembrokeshire. Field Studies, 1, (5) 1-31
30. NKLSON-SMITH, A (1964) "Some aspects of the Marine Ecology of Milford Haven; Pembrokeshire" Ph.D. thesis, University College of Swansea.
31. NHMSON-SMIIH, A. (1965). Marine Biology of Milford Haven: the physical environment, Field Studies, 2, 155-188.
32. NEHSON-SMITH, A. (1967). Marine Biology of Milford Haven: The distribution of Littoral Plants and Animals. Field Studies, 2 (4), 407-434.
33. NELSON-SMIIH, A. (1968a) "The effects of Oil pollution and Emulsifier cleansing on marine life in South West Britain, J. Applied Ecol. 5, 97-107
34. NELSON-SMITH, A. (1968b) The Biological consequences of oil pollution and shore cleansing, Field Studies, 2 (suppl) 73-80.
35. NEHSON-SMITH, A. (1970) The Problem of Oil Pollution of the Sea. Adv. Mar. Biol. 8
36. PARSON, R. (1972a) Some sub-lethal effects of refinery effluent upon the winkle Littorina sascatilis, O.P.R.U. Annual Report 1972, pp 21-3
37. PEIPIROON, S. (Unpublished) Ph.D. thesis, University of Wales, Swansea.
38. REISH, D.J. (1971) Effect of pollution abatement in Los Angeles harbours, Mar. Pollut. Bull. 2, 71-74
39. SANBORN, H. amd MALINS, D.C. (1977). Toxicity and metabolism of naphthalene: a study with marine larval invertebrates. Proc. Soc. Exp. Biol. Med.
40. SANDERS, H.L, GRASSLE, J.F. and HAMPSON, G.R. (1972) The West Falmouth oil spill, I. Biology, Reference No. 72-70, Woods Hole Oceanographic Institution, Woods Hole, Massachusetts.
41. SOKAL, R.R. and ROHLF, F.J. (1969) Biometry: The principles and practice of statistics in biological research.
42. STANDARD MEYHODS FOR THE BXAMINATION OF WATER AND HASTE WATER (1971) 13th Ed. American Public Health Association, New York. 254-256
43. TAIT, R.V. (1968) Elements of marine ecology. Butterworths
44. TEAL, J.M. (1076) Hydrocarbon uptake by deep sea benthes. pp 358-371 In Sources, Effects and Sinks of Hydrocarbons in the Aquatic Environment. All3S Washington, D.C.
45. VAN GELDER-OTHLAY; S. (1975) Some Physical and Biological Effects of Oil films floating on water, in Marine Ecology and Oil Pollution (ed. J.M. Baker)
46. WIILIAMSON, P. (1979) Opposite effects of age and weight in Cadmium concentrations of a gastropod mollusc. Ambio. 8 pp 30-31.
47. ZSOLNAY, A., MAYNARD, N.G. and GEBEREIN, C.D. (1977). Biogenic hydrocarbons in Intertidal communities: in Monitoring, and Enforcement: 1977 Oil Spill Conference, American Petroleum Institute.

